11 Icelandic summer spawning herring

11.1 Scientific data

11.1.1 Surveys description

The scientific data used for assessment of the Icelandic summer-spawning (ISS) herring stock derives from annual acoustic surveys (IS-Her-Aco-4Q/1Q), which have been ongoing since 1974 (Table 11.1.1.1). Normally these surveys are conducted in the period of October–January, but also as late as end of March. The surveyed area each year is decided on basis of available information on the distribution of the stock in the previous and the current year, which include information from the fishery. Thus, the survey area varies spatially as the survey is focused on the adult and incoming year classes but is considered to cover the whole stock each year.

The acoustic abundance index for the adult stock in the winter 2018/2019 derives from three dedicated acoustic surveys on RV Bjarni Sæmundsson (Óskarsson, 2019a): (1) A survey aiming at herring juveniles in the fjords northwest and north of Iceland in October; (2) A survey in December aiming at the fishable stock east and south of Iceland; (3). A survey in end of March aiming at the fishable stock at the main overwintering area of the stock west of Iceland.

In addition to getting an acoustic estimate on the adult part and on juveniles at age 1, the objective was also to get an estimate of prevalence of *Ichthyophonus* infection in the stock. The instrument and methods in the surveys were the same as in previous years and described in the stock annex and all the results are detailed in a WD to NWWG (Óskarsson, 2019a). The biological sampling in the survey is detailed in Table 11.1.1.2.

11.1.2 The survey results

The fishable part of the Icelandic summer-spawning herring stock was observed in two areas, west of Iceland in Kolluáll in end of March, and to a less extent southeast of Iceland in Breiðamerkurdjúp (Figure 11.1.2.1; Óskarsson, 2019a). The total acoustic estimate, according to these two surveys, came to 2.65 billion in numbers and the total biomass index was 350 kt (Table 11.1.1.1). The fishable part of the stock (≥ 27 cm) accounted for 42% in number and 94% of the biomass, or 328 kt. When only considering age 3+, the three most numerous year classes were those from 2013 (18%) and 2010 and 2011 (15 and 14%, respectively; Table 4). Like in last year, the 2014 year class was in relatively low number (6%). The total abundance index was lower than seen for years, or since the 1980s (Figure 11.1.2.2).

The annual survey aiming for the abundance of herring juveniles in the fjords northwest and north of Iceland took place in October 2018 (Óskarsson, 2019a). Fjords and areas covered (Figure 11.1.2.1) were comparable to other years. The juvenile survey is specially aimed for assessing the number-at-age 1. This because different from number-at-age 2, number-at-age 1 has been showed to give a signal of year class strength later at age 3 (Gudmundsdottir *et al.*, 2007). The herring juvenile surveys have been conducted in a comparable way since 1980, with several gaps in the time series. The survey in 2018 was considered to provide index applicable for stock assessment purpose. The abundance index of age 1 indicate that the 2017 year class is well above average size. Applying the linear-regression provided by Gudmundsdottir *et al.* (2007) implied that the 2017 year class will be 678 million at age 3 in the autumn 2020, while the average year class size across 1987–2014 is 585 million at age 3 and geometric mean of 518 million (ICES, 2018). This number should be used in the forecast in the 2019 assessment. The abundance index of age 1 in 2018 of 1.2 billion is lower than the highest indices in the time series from 1999 (2.8 billions), 1996 (1.6 billions) and 1991 (1.4 billion; Gudmundsdottir *et al.*, 2007), but equal to 2002 (1.2 billions) and above 1998 (1 billions), which both turned out to be strong year classes.

11.1.3 Prevalence of *Ichthyophonus* infection in the stock

A widespread ichthyophoniasis epizootic has been occurring in ISS-herring since 2008. This is caused by the parasite *Ichthyophonus* sp. Results of comprehensive analyses for the period 2008–2014 imply that significant infection mortality took place in the first three years after the outbreak started (2009–2011) but not the years after (2012–2016; Óskarsson *et al.*, 2018b). The level of the mortality was estimated with series of runs of the NFT-adapt assessment model, which gave the best fit to the data when applying infection mortality equivalent to 30% of the infected herring (heart inspection and survey abundance estimates provided $M_{infected}$) died annually in the first three years of the outbreak ($M_{year, age} = M_{fixed} + M_{infected, year, age} \times 0.3$; Table 11.3.2.1). M used in the stock assessment in 2018 for the year 2017 reflected these findings (ICES, 2018).

The prevalence of the *Ichthyophonus* infection in the stock in 2018/19 was estimated in a same way as has been done since the initiation of the infection in the autumn 2008 (Óskarsson and Pálsson, 2018). The prevalence of infection in 2018 was similar to 2017 for all year classes, except for the 2012 year class (Figure 11.1.3.1). It was also found to be high in the samples from the survey in March 2019 (26% vs. 32% in catch samples; Óskarsson 2019b). This is therefore an indication for new infection in the stock in 2018, even if at lower level than in 2017 and 2016 (Figure 11.1.3.1). This differs from the results obtained for the period 2010–2014, where analyses of younger age groups showed no indication of new infection, or at insignificant level. This pattern suggesting increased new infection in 2016–2018 can also be seen when the prevalence of infection by length is followed (Óskarsson, 2019b). Consequently, infection mortality is assumed to take place in 2019, like in 2016 and 2017. Thus, in the stock prognosis (Section 11.6), the abundance estimates from the final year of the assessment (1 January 2019) is lowered by this additional M as done in 2009–2011 assessments. The level of M should then follow the results by Óskarsson *et al.* (2018b), where age specific $M_{infected}$ (estimated from the catch samples; Figure 11.1.3.1) is multiplied by 0.3 and the fixed M (0.1) added to it. These M for 2019 (Table 11.3.2.1) should be used in the prognosis in 2019 and in the analytical assessment from 2020 and onwards, until better more reliable estimates become available.

11.2 Information from the commercial fishery

The total landings of ISS herring in 2018/2019 season was 40 683 t with no discards reported (Table 11.2.1 and in Figure 11.2.1). This includes also bycatches of ISSH in the mackerel and Norwegian spring-spawning herring (NSSH) fisheries in June–October 2018 (8.1 kt), where the part caught in June–August (6.1 kt) belongs actually to the official fishing season September 2017/August 2018. Including the summer catches in the subsequent fishing season, as done here, is a traditional handling of the catch data when assessing this stock. The quality of the herring landing data regarding discards and misreporting are consider adequate as implied in a general summary in Section 7 and in the Her-Vasu stock annex.

The recommended TAC for 2018/2019 fishing season (September–August; ICES, 2018) and allowable TAC (Regulation No. 633, 19 June 2018) was 35.186 kt (Table 11.2.1). Officially, according to the Directorate of Fisheries (http://www.fiskistofa.is/veidar/aflaupplysingar/heildar-aflamarksstada/), 1.7 kt remained unfished in March 2019, which needs to compensate the bycatch in the summer 2019.

The direct fishery started in end of October and lasted to December in offshore areas west of Iceland and contributed to 79% of the total catches (Figure 11.2.2). The remaining 21% (8.1 kt) of the catch was taken as bycatch in the fishery for NSSH and mackerel during June to October and mainly southeast and east of Iceland (Figure 11.2.2).

Spring-spawning herring (assumed to be Icelandic spring spawners and not NSSH) was mixed with the Icelandic summer-spawning herring stock in the catches in the autumn 2018 as normally observed (Óskarsson, 2018b). The proportion became to 1.3% for the areas west of Iceland during October–December. This proportion is a bit lower than was observed for the autumns 2013–2017 (2–5%; Óskarsson, 2018b).

11.2.1 Fleets and fishing grounds

The herring fishing season has taken minor changes in the last three decades as detailed in the stock annex. All seasonal restricted landings, catches and recommended TACs since 1984 are given in thousands tonnes (kt) in Table 11.2.1.

All the catch in 2018/2019 was taken in pelagic trawls (Figure 11.2.1), which reflects that both the targeting and bycatch fisheries takes mainly place in offshore areas. During all fishing seasons from 2007/2008 to 2012/2013, most of the catches (~90%) were been taken in inshore areas west off Iceland in Breiðafjörður, while prior to that they were mainly taken off the south-, southeast-, and the east coast. In 2013/2014 there was an indication for changes in this pattern, with less proportion in Breiðafjörður, and then in 2014/2015 almost all of the overwintering west of Iceland took place offshore, which continued this winter. These changes in the stock distribution explain the dominance of pelagic trawl in the fishery, which is preferred by the fleet over purse seine in offshore areas.

To protect juveniles herring (27 cm and smaller) in the fishery, area closures are enforced based on a regulation of the herring fishery set by the Icelandic Ministry of Fisheries (no. 376, 8 October 1992). No closure was enforced in this herring fishery in 2017/18. Normally, the age of first recruitment to the fishery is age-3, which is fish at length around 26–29 cm.

11.2.2 Catch in numbers, weight at age and maturity

Catch at age in 2018/2019:

The procedure for the catch at age estimations, as described in the Stock Annex, was followed for the 2018/19 fishing season. It involves calculations from catch data collected at the harbours by the research personnel (0%) or at sea by fishermen (100%). This year, the calculations were accomplished by dividing the total catch into four cells confined by season and area as detailed in Óskarsson (2019b). In the same way, three weight-at-length relationships derived from the length and weight measurements of the catch samples were used. On basis of difference in length-at-age between the summer months (July–September) and the winter (October–December), two length-age keys were applied. The catches of the Icelandic summer spawners in number-at-age for this fishing season as well as back to 1982 are given in Table 11.2.2.1. The geographical location of the sampling is shown on Figure 11.2.2.

The age composition in the direct winter fishery 2018/2019 was different from the composition in the bycatch of herring in the mackerel and NSS-herring fishery in the summer 2018 (Figure 11.2.2.1). The summer fishery included to a higher degree younger age groups (e.g. age 3-4 contributed to ~33% of the biomass) than the direct fishery in the west (13%), and consequently vice versa for older age groups. This difference is reflecting the geographical distribution of the different age groups, with higher proportion of younger age groups in the east and south than in the west. Thus, the size of these two relatively small year classes, especially the one at age 3 (2015 y.c.), are likely to be somewhat underestimated in the assessment. Their appearance in the acoustic measurements (Figure 11.2.2.1; Óskarsson, 2019a) will though reduce this chance of underestimation.

Weight at age:

As stated in the stock annex, the mean weight-at-age of the stock is derived from the catch samples (Table 11.2.2.2). The total number of fish weighed from the catch in 2018/19 was 1060 (24 samples) and 985 of them were aged from their fish scales.

Proportion mature:

The fixed maturity ogives were used in this year's assessment, as described in detail in the stock annex, where proportion mature-at-age 3 is set 20% and 85% for fish at age 4, while all older fish is considered mature.

Observed versus predictions of catch composition:

The relative contribution of the different year classes was somewhat different from what was predicted in the analytical assessment in 2018 (Figure 11.2.2.1), and it applies particularly to the 2013 and 2014 year classes, which were more numerous than predicted in contrast to older age groups (age 9+).

11.2.3 Error in the assessment and advice in 2018 and 2017

In the preparation for the 2019 assessment, an error was observed in the tuning series for the autumn 2016. It was forgotten to shift the age when representing 1 January 2017, meaning that number-at-age 3 (etc) in the autumn 2016 was set to age 3 in 2017 instead of age 4. This error influenced both the 2017 and 2018 assessments. Its impacts was verified for the 2018 assessment and advice and showed that the reference biomass of age 4+ in the beginning of 2018 was 250.7 kt instead of 234.6 kt, which should have resulted in advice of 37 604 tonnes instead of advised TAC of 35 186 tonnes (Figure 11.2.3.1). The advice should have been 2418 tonnes higher or 6.9% higher. The WG considered this difference to be minor and irrelevant when considering the assessment quality and status of the stock. Besides, the targeted fishery in the 2018/2019 season has already taken place. Thus, the advice from 2018 should not be revised but this error in the time series has of course been corrected in this year's assessment.

11.3 Analytical assessment

11.3.1 Analysis of input data

Examination of catch curves for the year classes from 1987 to 2014 (Figure 11.3.1.1) indicates, in general, that the total mortality signal (Z) in the fully recruited age groups is around 0.4. It is under the assumption that the effort has been the same the whole time. In recent years the effort has changed a lot because of the infection and spatial distribution of the stock, and the mass mortality in 2012/2013, which makes any strong deductions from the catch curves for those recent less meaningful.

Catch curves were also plotted using the age disaggregated survey indices for each year class from 1987–2014 (Figure 11.3.1.2). Even if the total mortalities look at bit noisy for some year classes, they seem to be fairly close to 0.4, for example for 1996–2008 year classes. There is an indication that the fish is fully assessable to the survey at age 3–5.

Increased mortality in the stock because of the *Ichthyophonus* outbreak cannot be detected clearly from the catch curves of the surveys. However, considering that F was reduced drastically in the beginning of the outbreak, similar Z means an increased M during that period, representing infection mortality.

11.3.2 Exploration of different assessment models

Input data:

In order to explore the data this year, two models were run, NFT-ADAPT (VPA/ADPAT version 3.3.0 NOAA Fisheries Toolbox) that has been used as the basis for the assessments since 2005 and a separable model (Muppet) also used in the MSE in 2017 for the stock (ICES 2017b; Björnsson 2018) as well as analytical assessment of Icelandic saithe (ICES, 2018). Applying NFT-ADAPT was evaluated at benchmark assessment in January 2011 (ICES, 2011a) and it found to be appropriate as the principal assessment tool for the stock. The catch data used were from 1987/88–2018/19 (Table 11.2.2.1) and survey data from 1987/88–2018/19 (Table 11.1.1.1). Other input data consisted of: (i) mean weight at age (Table 11.2.2.2); (ii) maturity ogive (Table 11.2.2.3); (iii) natural mortality, M, that was set to 0.1 for all age groups in all years, except for 2009–2011 and 2017–2018 where additional age dependent mortality was applied because of the *Ichthyophonus* infection (see Section 11.1.3; Table 11.3.2.1; Óskarsson *et al.*, 2018b); (iv) proportion of M before spawning was set to 0.5; and (v) proportion of F before spawning was set to 0. Thus, in comparison to last year's assessment, all the input data are the same with an additional year of data.

Results:

The estimated parameters in NFT Adapt are the stock in numbers at age. The parameters are output by the Levenburg-Marquardt Non-Linear Least Squares minimization algorithm (see VPA/ADAPT Version 3.3.0, Reference Manual). The estimated parameters were stock numbers for ages 4 to 12 in the beginning of year 2018, while the stock numbers at age 3 was derived from survey estimates in 2017 (i.e. projection from age-1 survey index to age-3 according to Gudmundsdóttir *et al.,* 2007 and recommended by ICES (2011a)) instead of geometric mean as default in the model. Like in last years' assessments, the *input partial recruitment* was set to 1 for ages 4 and older and the *classic* method was used to calculate the value of fully-recruited fishing mortality in the terminal year.

The catchability at age in the survey, as estimated by the NFT Adapt, and the CV is shown in Figure 11.3.2.1. The age groups 3–10 were used for tuning (Table 11.1.1.1 as decided at the benchmark in ICES (2011a). In comparison to last year, the catchability of the survey is relatively the same with similar uncertainty.

The output and model settings of the NFT-Adapt run (the adopted final assessment model) are shown in Table 11.3.2.2. Stock numbers and fishing mortalities derived from the run are shown in Table 11.3.2.3 and Table 11.3.2.4, respectively, and summarized in Table 11.3.2.5 and Figure 11.3.2.2.

Residuals of the model fit are shown in Figure 11.3.2.3 and Table 11.3.2.6, and shows both cohort and year affects. The main pattern is the same as presented in recent assessments. Positive residuals, where the model estimates are smaller than seen in the survey, can be seen for 1994 and 1999 year classes for almost all age groups and a negative residuals for the 2001 and 2003 year classes. Year blocks of positive residuals are apparent for the years ~2000 to 2006 (i.e. referring to 1 January). During these years, the stock was overwintering in offshore areas off the east and west coast, compare to mainly easterly distribution before and overwintering in inshore areas there after (from ~2006–2012). These positive blocks could therefore reflect changes in catchability of the survey for these years. After 2008 the residuals are generally behaving well.

Retrospective analyses indicate a consistency over the most recent six years, i.e. adding new data to the model does not change the present perception of the stock size much (Figure 11.3.2.4). The small upward revision for the last years is likely caused by the increased M in 2017 and 2018 (due to infection mortality), and for compensating for it, the model increased the stock size back in time. This is a pattern seen before (ICES, 2017c). The retros for the fishing mortality and recruits behave, in a same way, well for the last four years.

Like demonstrated and analysed earlier (ICES, 2014), the main difference between observed and predicted survey values from the NFT-Adapt model was for the period 1999–2004, where the observed values were well above the predicted (Figure 11.3.2.5), otherwise they fitted relatively well. Like seen in the residual plot (Figure 11.3.2.3), the observed value for the 2009 survey was lower than predicted and the vice versa for the 2012 survey (referring to the beginning of the year; Figure 11.3.2.5). The low survey value in 2009 is likely underestimate due to distribution of the stock that year in the fjord west of Iceland (Breiðafjörður; Óskarsson *et al.*, 2010), while the positive block during 2000–2004 was previously found to be mainly caused by the large 1999 year class (ICES, 2014) and possibly changes in the catchability of the survey as suggested above. However, an exploratory run in NFT-Adapt done in the 2011 assessment (ICES, 2011b) where these years were excluded in the tuning, did not change the point estimate of the stock size in the latest year (1 January 2011), implying that the terminal point estimates in the final run was not driven by this residual block.

F in 2018 for age 4 (0.48) and age 5 (0.29) was much higher than the $F_{Avg.5\text{-}10}$ and $F_{W5\text{-}10}$ (0.166 and 0.175, respectively) despite their low catches (Table 11.3.2.4). This is related to that these year classes were mainly caught during the summer fishery 2018 and hardly observed in the catches in the direct fishery west of Iceland 2017/18 and were, consequently, assessed small by the NFT-Adapt model. In other words, this is a consequence of distribution of these age groups and not being targeted in the main fishery in 2018/2019, which might result in underestimation of their size. This adds uncertainty to the assessment and the stock projection (see Section 11.6).

Comparisons of different models:

The two models explored, NFT-Adapt and the separable model (Muppet), gave very similar results, and especially for the latest years of the assessments (Figure 11.3.2.2). This indicates that the results are driven by the input data and not by the model used.

11.3.3 Final assessment and TAC advice on basis of Management Plan

This is an update assessment so the results of the NFT-Adapt were adopted as point estimator for the prediction and thus the basis for the advice as in recent years. The model settings and outputs are shown in Table 11.3.2.2 to Table 11.3.2.4 and Figure 11.3.2.2.

The final assessment indicates that the reference biomass of age 4+ in the beginning of the year 2019 is 196.588 kt. SSB in 2019 will be 197.976 kt or slightly below MGT $B_{trigger}$ = B_{lim} =200 kt, meaning that the slighting rule applies. Thus, the $TAC_{2019/20}$ according to the Management Plan (Section 11.4) is 0.15* (198/200) * (196.588 kt) = 29.190 kt.

11.4 Reference points and the Management plan

Precautionary approach reference points:

The working group points out that managing this stock at an exploitation rate at or above $F_{0.1}$ = F_{MSY} = 0.22 has been successful in the past for almost 30 years, despite biased assessments. At the 2016 NWWG meeting, the PA reference points for the stock were verified and revised (ICES, 2016). On basis of the stock-recruitment relationship deriving from time-series ranging from 1947–2015, keeping B_{lim} = 200 kt was considered reasonable as the Study Group on Precautionary Reference Points for Advice on Fishery Management concluded also in February 2003. Other PA reference points were derived from B_{lim} and these data in accordance to the ICES Advice Technical Guidelines and became these: B_{pa} = 273 kt ($B_{pa} = B_{lim} \times e^{1.645\sigma}$, where $\sigma = 0.19$); F_{lim} = 0.61 (F that leads to SSB = B_{lim}, given mean recruitment); F_{pa}= 0.43 ($F_{pa} = F_{lim} \times \exp(-1.645 \times \sigma)$, where $\sigma = 0.18$).

MSY based reference points:

At a NWWG meeting in 2011 an exploratory work, using the HCS program Version 10.3 (Skagen, 2012), was used to evaluate possible points based on the MSY framework that could be a basis for a management plan and Harvest Control Rule later (ICES, 2011b). Number of different runs was made with varying settings. The results implied that the MSY framework was confirmative with the currently used precautionary reference points. It means that the currently used $F_{0.1}$ = 0.22 could be a valid candidate for F_{MSY}. During a Management Strategy Evaluation (MSE) for the stock in April 2017 (ICES, 2017b), F_{MSY} = 0.22 was not considered to be significantly different from results of simulation giving 0.24. Thus, it was concluded adequate to keep F_{MSY} = 0.22.

Management plan

A Management Strategy Evaluation (MSE) for the stock took place in 2017 (ICES, 2017b). Five different HCRs were tested and all of them, except for the advisory rule applied at that time (F_{MGT} = 0.22), were considered precautionary and in accordance with the ICES MSY approach. One of these HCR was later adopted by Icelandic Government as a Management plan for the stock. This HCR is based on reference biomass of age 4+ in the beginning of the assessment years ($B_{ref,\,Y}$), a spawning stock biomass trigger (MGT $B_{trigger}$) is defined as 200 kt, and the harvest rate (HR_{MGT}) is set as 15% of the reference biomass age4+ in the beginning of the assessment year. In the assessment year (Y) the TAC in the next fishing year (1 September of year Y to 31 August of year Y+1) is calculated as follows:

When SSB_Y is equal or above MGT $B_{trigger}$:

$TAC_{Y/y+1} = HR_{MGT} * B_{Ref,y}$

When SSB_Y is below MGT $B_{trigger}$:

$TAC_{Y/y+1} = HR_{MGT} * (SSB_y / MGT\ B_{trigger}) * B_{ref,y}$

In the MSE simulation, the ongoing *Ichthyophonus* epidemic was considered to continue and was accounted for. Consequently, this HCR is independent of estimated level of *Ichthyophonus* mortality and requires no further action during such epidemics.

The distribution of the realized harvest rate when the HCR is followed showed that the 90% expected range are within a harvest rate of 0.099–0.22 with no bias and 0.122–0.247 if bias is applied. The recent realized harvest rates are within the above range

11.5 State of the stock

The stock was at high levels until around late 2000s but since then a substantial reduction has taken place despite a low fishing mortality. The reduction is consequence of mortality induced by *Ichthyophonus* outbreak in the stock in 2009–2011 and 2016–2018 in addition to small year classes entering the stock since around 2005, particularly the 2011–2014 year classes. Hence, SSB will be below MSY $B_{trigger}$ in 2019 but above the MSG $B_{trigger}$ and B_{lim}. Survey indices from the autumn 2018 indicate that the 2017 year class might be well above average but it will enter the fishable stock first in 2021.

11.6 Short term forecast

11.6.1 The input data

The final adopted model, NFT-Adapt, which gave the number-at-age on 1 January, 2019, was used for the prognosis. All input values for the prognosis are given in Table 11.6.1.1. Because of the expected *Ichthyophonus* mortality in the stock in the spring 2019 (see Section 11.1.3), the NFT-Adapt model output were reduced according to the infection ratios times 0.3 (Table 11.3.2.1), or the same approach as used in the assessments in 2009–2011 and 2018 (ICES, 2011b; 2018a; Óskarsson *et al.*, 2018b).

The weights were estimated from the last year catch weights (see Stock Annex) and as in the recent years, the weights are expected to continue to be high, except for the youngest age groups, which is though still well within observed range (Figure 11.6.1.1). The weight for age 3 was set equal to the value used in 2018 (ICES 2018) because the weight deriving from the formula provided in the Stock Annex gave much lower and unrealistic value.

According to the Stock Annex, the selection pattern in the prognosis should be based averages over 2016 to 2018 from the final run. Because of the high F for age 4 in 2017 and 2018 (Section 11.3.2), the expected selection pattern for them in 2019 became unrealistically high (Figure 11.6.1.2), and thereby much higher than in recent years. Consequently, it was decided to deviate from the Stock Annex and base the selection only on the 2016 value. Similar thing was done in the 2018 assessment (ICES, 2018) and is justified by the fact that it gives a more realistic estimate and that the advice deriving from the assessment is not based on the outcome of this forecast. This high selection pattern for age 4 in 2017 and 2018 is expected to be related to likely underestimation of abundance of these age groups, whereas they are in small number in the catches as they are not found in the main fishing grounds west of Iceland and appear almost exclusively in the bycatch during the summers (see Section 11.2).

As traditionally, M was set 0.1, proportion M before spawning was set 0.5 and proportion F before spawning was set 0. The numbers of recruits in the prognosis were determined as follows:

The 2016 year class: The planned acoustic survey in the autumn 2017 aimed for getting an abundance index for this year class was cancelled due to vessel problems. An abundance index for this year class derives from a survey in February 2018 where a single fjord was covered (Eyjafjörður) and the assumed adult areas east and south of Iceland (Óskarsson, 2018a). This index was used by convert it to number at age 3 in accordance to Gudmundsdóttir *et al*. (2007) that provides estimate of 360 million at age 3 in 2019.

The 2017 year class: An acoustic survey aimed for getting an abundance index for this year class took place in September–October 2018 (Óskarsson, 2018a), and convert it to number at age 3 in accordance to Gudmundsdóttir *et al*. (2007) provides estimate of 678 million at age 3 in 2020.

In summary, the basis for the stock projection is as follows: SSB(2019) = 212 kt; Biomass age 4+ (1 January 2019) = 231 kt; Catch (2018/19) = 40 kt; $WF_{5\text{–}10}$(2018) = 0.175; HCR (2018) = 0.145. There are deviations from the Stock Annex (weight of age 3 and selection of age 4), which are considered to be of improvements and not of concern since this projection has no impacts on the advice.

11.6.2 Prognosis results

SSB in the beginning of the fishing season 2019/20 (approximately the same time as spawning in July 2019) is estimated to be 212 kt, which is above MGT $B_{trigger}$ of 200 kt. Consequently, advised TAC on basis of the Management rule is 0.15 × Biomass 4+ (230.48 kt) = 34.572 kt. This results in $F_{W5\text{–}10}$ = 0.175 in 2019/20 and SSB = 221 kt in 2020 (Table 11.6.2.1).

The results of different options are given in Table 11.6.2.1. The catch in the season 2019/20 is projected to come from most age groups in equal portions (8–13%) except for year classes from 2014 and 2017 (3% each; Figure 11.6.2.1).

11.7 Medium term predictions

Because of the increased uncertainty of the assessment in relation to the development of the *Ichthyophonus* outbreak in the coming months and years, the uncertainty in size of the recruiting year classes, and the new management rule, no medium-term prediction is provided.

11.8 Uncertainties in assessment and forecast

11.8.1 Uncertainty in assessment

There are number of factors that could lead to uncertainty in the assessment. Two of them are addressed here. Additional natural mortality caused by the *Ichthyophonus* infection was set for the first three years of the outbreak (2009–2011) and in 2017 and 2018 ($M_{infected, age, year}$ multiplied by 0.3 (see Section 11.1.3). This quantification of the infection mortality based on Óskarsson *et al.* (2018b), was considered to improve the assessment and reduce its uncertainty. For the most recent years, where new infection reappeared (2017–2019), more accurate estimation of the infection mortality will be possible in the years to come but until then, this approach will add uncertainty to the assessment. Worth noticing, increasing M has been shown to increase the historical perception of the stocks size but has minor impacts on the assessment of the final year and the resulting advice.

The signals from the last catches and the surveys give somewhat contradicting results about the size of the 2013–2015 year classes (Figure 11.2.2.1), even if all of them appears to be small, particularly the 2014 year class. The size of these year classes is therefore not very well determined yet, which adds uncertainty to the assessment. Considering that the direct winter fishery west of Iceland is not targeting these year classes, which are mainly found southeast and east of Iceland, their size is more likely to be underestimated in the analytical assessment.

11.8.2 Uncertainty in forecast

It is important to notice that the advice for 2019/2020 fishing season deriving from the Management plan is independent of the forecast and its uncertainty as it is only based on the reference biomass in the beginning of the assessment year. The uncertainty in the assessment mentioned above related to the apparent new infection in the stock in 2017–2019 and size of the recruiting year classes, apply also for the forecast.

Moreover, the number-at-age 3 in the beginning of the year 2019 used in the prognosis (360 millions) was predicted from a survey estimate of number at age 1 in 2017 in accordance with the approach described in the Stock Annex. This index derives from an incomplete survey but is used here as it is in accordance with the Stock Annex, the survey covered the single most important nursery grounds of the stock (Eyjafjörður), it was considered to be more appropriate than applying geometric mean, and this decision has no impact on the fishing advice. Thus the resulting stock size in 2020 is likely to be too pessimistic.

11.8.3 Assessment quality

For a period, there was concerns regarding the assessment because of retrospective patterns of the results. No assessment was provided in the 2005 due to data and model problems and in the two next consecutive years, ACFM rejected the assessment due to the retrospective pattern. In the assessments in 2007–2009 there was observed an improvement in the pattern from NFT-Adapt, while in 2010–2011, a retrospective pattern appeared again which was both related to the high M because of the *Ichthyophonus* infection but also due to new and more optimistic information about incoming year classes to the fishable stock (particularly the 2008 year class) and fishing pattern in recent year. The retrospective pattern in the last five and this year's assessment are less than seen for many years for SSB and F (Figure 11.3.2.4). Simultaneously the residuals from the survey are behaving better than before (Figure 11.3.2.3). This together could be interpreted as indications for improvements in the assessment quality in recent years in comparison to the years before. The small retros in the SSB for this year's assessment is considered to be related to the additional infection mortality set for 2017–2018, where the model increase the stock size back in time to compensate for the increase M.

As stated in the 2017 NWWG report (ICES, 2017c), the revision of the infection mortality applied in the analytical assessment for the years 2009–2011 in accordance to the estimated mortality levels (Section 11.1.3), is also considered as an improvement of the assessment. Thus, the downward revision of the stock size over the period ~2003–2011 compared to the last year's assessment (Figure 11.3.2.2) is considered to provide more robust figure of development in the historical stock's size.

11.9 Comparison with previous assessment and forecast

This year's assessment was conducted in the same way as in last year, apart from the correction on the survey indices from 2017 (see Section 11.2.3). Additional natural mortality was applied to 2017–2018 because of the infection (see Section 11.1.3), which caused an upward revision of the stock size for the most recent years (Figure 11.3.2.4). When the estimates for 1 January 2018 are compared with last year's assessment, the results of the final NFT run in 2019 gives a more optimistic view on the size of the small 2013 and 2014 year classes (Figure 11.3.2.6). Apart from that there is not a big difference. Note that the estimate of the 2015 year class in 2018 was based on a survey estimation while in the assessment model in 2019.

11.10 Management consideration

Inspections indicate still a high prevalence of heart lesions related to *Ichthyophonus hoferi* in the herring stock. More importantly, new infection has been taken place in the stock last three winters but possibly with a decreased intensity in 2018/2019. Significant new infection was otherwise last observed in 2010 (Óskarsson *et al.*, 2018b). Correspondingly, induced mortality due to the infection was unavoidably applied for 2017–2019, and this second outbreak might continue in the coming year. Considering the presently low stock size, the ongoing second outbreak, and continuing poor year classes entering the fishable stock, the stock size will most likely remain at low level in the next two years and be between B_{lim} and MSY $B_{trigger}$. The survey results implying large 2017 year class might change this situation from 2021 onwards when it starts to enter the fishable stock.

11.11 Ecosystem considerations

The reason for the outbreak of *Ichthyophonus* infection in the herring stock that was first observed in the autumn 2008 is not known but is probably the effect of interaction between environmental factors and distribution of the stock (Óskarsson *et al.* 2009). It includes that outbreak of *Ichthyophonus* spores in the environment, which infect the herring via oral intake (Jones and Dawe, 2002), could be linked to the observed increased temperature off the southwest coast. Further researches on the causes and origins of such an outbreak are ongoing at MFRI. It involves scanning for *Ichthyophonus* DNA in zooplankton species that the herring feeds on with PCR (Polymerase chain reaction) technique. Results from that work (MS thesis) can be expected in the summer 2019, while preliminary results indicate that the source of the infection is widespread and is in various zooplankton groups and species. With respect to the impacts of the outbreak on the herring stock, recent analyses show that significant additional mortality took place over the first three years only (Óskarsson *et al.*, 2018b), despite a high prevalence of infection for now nine years. As pointed out above, a new infection since the summer 2016 is however, expected to cause significant mortality again. For how long time this outbreak will last is unknown as this is basically an unprecedented outbreak. The signs of the infection that is found in the stock will most likely remain for some years, even if no new infection will occur, and then decrease and disappear over some years as new year classes replace the older ones. The observed new infection will however delay this process.

All general ecosystem consideration with respect to the stock can be found in the Ecosystem Overview for the Icelandic Ecoregion (ICES, 2017a).

11.12 Regulations and their effects

The fishery of the Icelandic summer-spawning herring is limited to the period 1 September to 1 May each season, according to regulations set by the Icelandic Fishery Ministry (**no. 770, 8 September 2006**). Several other regulations are enforced by the Ministry that effect the herring fishery. They involve protections of juvenile herring (27 cm and smaller) in the fishery where area closures are enforced if the proportion of juveniles exceeds 25% in number (no. 376, 8 October 1992). No such closures took place in 2017/2018. Another regulation deals with the quantity of bycatch allowed. Then there is a regulation that prohibits use of pelagic trawls within the 12 nautical miles fishing zone (**no. 770, 8 September 2006**), which is enforced to limit bycatch of juveniles of other fish species.

11.13 Changes in fishing technology and fishing patterns

There are no recent changes in fishing technology which may lead to different catch compositions. The fishing pattern in the seasons 2014/2015 to 2018/2019 was different from the previous seven seasons. Instead of fishing near only in a small inshore area off the west coast in purse seine, the whole directed fishery took place in offshore areas west of Iceland by pelagic trawls. These changes are not considered to affect the selectivity of the fishery because the fishery is still targeting dense schools of overwintering herring in large fishing gears, getting huge catches in each haul and is by none means size selective.

Bycatch of Icelandic summer-spawning herring in summer fishery for NE-Atlantic mackerel and Norwegian spring-spawning herring has been taken place since around mid-2000s. Until that time, no summer fishery on this stock had taken place for decades. Part of this bycatch is on the stock components (e.g. juveniles and herring east of Iceland) that are not fished in the direct fishery on the overwintering grounds in the west. However, these bycatches are well sampled and contributes normally to less than 10% of the total annual catch, but were as high as 37% in the season 2017/2018. It can be explained by the low TAC, so the fleet did not have much quota left for direct autumn fishery. Still, the impacts of these changes on the assessment are considered to be insignificant.

The fishing pattern varies annually as noted in Section 11.2 and it is related to variation in winter distribution of the different age classes of the stock. This variation can have consequences for the catch composition but it is impossible to provide a forecast about this variation.

11.14 Species interaction effects and ecosystem drivers

The WG have not dealt with this issue in a thoroughly and dedicated manner. However, some work has been done in this field in recent years in one way or another.

Regarding relevant researches on species interaction, the main work relates to the increasing amount of North East Atlantic mackerel (NEAM) feeding in Icelandic waters after 2006 (Astthorsson *et al.*, 2012; Nøttestad *et al.*, 2016). Surveys in the summers since 2010 indicate a high overlap in spatial and temporal distribution of NEAM and Icelandic summer-spawning herring (Óskarsson *et al.*, 2016). Moreover, the diet composition of NEAM in Icelandic waters showed a clear overlap with those of the two herring stocks, i.e. Icelandic summer-spawning herring and Norwegian spring-spawning herring (Óskarsson *et al.*, 2016). Even if copepoda was important diet group for all the three stocks its relative contribution to the total diet was apparently higher for NEAM than the two herring stocks. Considering former studies of herring diet, this finding was unexpected, and particularly how little the copepoda contributed to the herring diet. This difference in the stomach content of NEAM and the two herring stocks indicated that there could be some difference in feeding ecology between them in Icelandic waters, where NEAM preferred copepoda, or feed in the water column where they dominate over other prey groups, while the opposite would be for the herring and the prey Euphausiacea. Recent studies in the Nordic Seas have shown similar results (Langøy *et al.*, 2012; Debes *et al.*, 2012). The indication for difference in feeding ecology of the species is further supported by the fact that the body condition of the two herring stocks showed no clear decreasing trend since the invasion of NEAM started into Icelandic waters. On the contrary the mean weights-at-age (and at-length) of the summer spawners have been high after 2010 (Óskarsson, 2019b) and for example record high in the autumn 2014 (Figure 11.6.1.1). It should though be noted that comparison of the diet composition of herring in recent years to earlier studies, mainly on NSS herring, indicate that the herring might have shifted their feeding preference towards Euphausiacea instead of Copepoda. That is possibly a consequence of increased competition for food with NEAM, where the herring is overwhelmed and shifts towards other preys.

The WG is not aware of documentations of strong signals from ecosystem or environmental variables that impact the herring stock and could possibly be a basis for implementing ecosystem drivers in the analytical basis for its advice. For example, recruitment in the stock has been positively, but weakly, linked to NAO winter index (North Atlantic Oscillation) and sea temperature (Óskarsson and Taggart, 2010), while indices representing zooplankton abundance in the spring have not been found to impact the recruitment (Óskarsson and Taggart, 2010) or body condition and growth rate of the adult part of the stock (Óskarsson, 2008). Considering these relations derived from the historical data, relatively warm waters around Icelandic (MRI 2016), and high positive NAO in recent years (http://www.cpc.ncep.noaa.gov/products/precip/CWlink/pna/nao.shtml), it was concluded in last year's report (ICES, 2018) that we could expect a good recruitment in the stock. It seems to be coming about with an encouraging first measurement of the 2017 year class.

11.15 Comments on the PA reference points

The WG dealt with the reference points in 2016 and revised them in accordance to the ICES Technical Guidelines (ICES, 2016).

11.16 Comments on the assessment

The assessment implies that the stock size has been declining since end of 2000s due to a combination of *Ichthyophonus* mortality and series of below average and poor year classes entering the stock. The 2014 year class entering the reference biomass and SSB in 2018 is estimated record small despite an upward revision from last year's assessment. However, its size is yet poorly determined, and the assessment might therefore still be rather pessimistic.

There is compelling evidence for new infection by *Ichthyophonus* in the stock in the winter 2018/19, even if less intensive than in the two years before. This called for applying additional infection mortality in 2019 until spawning. This decision has no impacts on the advice based on the management plan, but lowered the SSB in 2019. The mortality level for 2017–2019 cannot be estimated adequately with data at hand but can within several years. This current outbreak adds uncertainty to the assessment and advice.

An error in the 2017 and 2018 assessments was observed in 2019 (Section 11.2.3). This error had minor impacts on the assessments and was not considered to deserve a revision of the advice from them.

11.17 References

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11.18 Tables

Table 11.1.1.1. Icelandic summer-spawning herring. Acoustic estimates (in millions) in the winters 1973/74–2018/19 (age refers to the autumns). No surveys (and gaps in the time-series) were in 1976/77, 1982/83, 1986/87, 1994/95.

Year\age	2	3	4	5	6	7	8	9	10	11	12	13	14	15+	Total
1973/74	154.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	154
1974/75	5.000	137.000	19.000	21.000	2.000	2.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	186
1975/76	136.000	20.000	133.000	17.000	10.000	3.000	3.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	322
1977/78	212.000	424.000	46.000	19.000	139.000	18.000	18.000	10.000	0.000	0.000	0.000	0.000	0.000	0.000	886
1978/79	158.000	334.000	215.000	49.000	20.000	111.000	30.000	30.000	20.000	0.000	0.000	0.000	0.000	0.000	967
1979/80	19.000	177.000	360.000	253.000	51.000	41.000	93.000	10.000	0.000	0.000	0.000	0.000	0.000	0.000	1004
1980/81	361.000	462.000	85.000	170.000	182.000	33.000	29.000	58.000	10.000	0.000	0.000	0.000	0.000	0.000	1390
1981/82	17.000	75.000	159.000	42.000	123.000	162.000	24.000	8.000	46.000	10.000	0.000	0.000	0.000	0.000	666
1983/84	171.000	310.000	724.000	80.000	39.000	15.000	27.000	26.000	10.000	5.000	12.000	0.000	0.000	0.000	1419
1984/85	28.000	67.000	56.000	360.000	65.000	32.000	16.000	17.000	18.000	9.000	7.000	4.000	5.000	5.000	689
1985/86	652.000	208.000	110.000	86.000	425.000	67.000	41.000	17.000	27.000	26.000	16.000	6.000	6.000	1.000	1688
1987/88	115.544	401.246	858.012	308.065	57.103	32.532	70.426	36.713	23.586	18.401	24.278	10.127	3.926	4.858	1965
1988/89	635.675	201.284	232.808	381.417	188.456	46.448	25.798	32.819	17.439	10.373	9.081	5.419	3.128	5.007	1795
1989/90	138.780	655.361	179.364	278.836	592.982	179.665	22.182	21.768	13.080	9.941	1.989	0.000	0.000	0.000	2094
1990/91	403.661	132.235	258.591	94.373	191.054	514.403	79.353	37.618	9.394	12.636	0.000	0.000	0.000	0.000	1733
1991/92	598.157	1049.990	354.521	319.866	89.825	138.333	256.921	21.290	9.866	0.000	9.327	0.000	0.000	1.494	2850
1992/93	267.862	830.608	729.556	158.778	130.781	54.156	96.330	96.649	24.542	1.130	1.130	3.390	0.000	0.000	2395
1993/94	302.075	505.279	882.868	496.297	66.963	58.295	106.172	48.874	36.201	0.000	4.224	18.080	0.000	0.000	2525
1995/96	216.991	133.810	761.581	277.893	385.027	176.906	98.150	48.503	16.226	29.390	47.945	4.476	0.000	0.000	2197
1996/97	33.363	270.706	133.667	468.678	269.888	325.664	217.421	92.979	55.494	39.048	30.028	53.216	18.838	12.612	2022
1997/98	291.884	601.783	81.055	57.366	287.046	155.998	203.382	105.730	35.469	27.373	14.234	36.500	14.235	11.570	1924
1998/99	100.426	255.937	1081.504	103.344	51.786	135.246	70.514	101.626	53.935	17.414	13.636	2.642	4.209	8.775	2001
1999/00	516.153	839.491	239.064	605.858	88.214	43.353	165.716	89.916	121.345	77.600	21.542	3.740	11.149	0.000	2823
2000/01	190.281	966.960	1316.413	191.001	482.418	34.377	15.727	37.940	14.320	15.413	14.668	1.705	3.259	0.000	3284
2001/02	1047.643	287.004	217.441	260.497	161.049	345.852	62.451	57.105	38.405	46.044	38.114	21.062	3.663	0.000	2586
2002/03	1731.809	1919.368	553.149	205.656	262.362	153.037	276.199	99.206	47.621	55.126	18.798	24.419	24.112	1.377	5372
2003/04	1115.255	1434.976	2058.222	330.800	109.146	100.785	38.693	45.582	7.039	6.362	7.509	10.894	0.000	2.289	5268
2004/05	2417.128	713.730	1022.326	1046.657	171.326	62.429	44.313	10.947	23.942	12.669	0.000	1.948	11.088	0.000	5539
2005/06	469.532	443.877	344.983	818.738	1220.902	281.448	122.183	129.588	73.339	65.287	10.115	9.205	3.548	12.417	4005
2006/07	109.959	608.205	1059.597	410.145	424.525	693.423	95.997	123.748	48.773	0.955	0.000	0.000	0.000	0.480	3576
2007/08	90.231	456.773	289.260	541.585	309.443	402.889	702.708	221.626	244.772	13.997	22.113	68.105	10.136	2.800	3376
2008/09	149.466	196.127	416.862	288.156	457.659	266.975	225.747	168.960	29.922	26.281	17.790	9.881	0.974	3.195	2258

Year\age	2	3	4	5	6	7	8	9	10	11	12	13	14	15+	Total
2009/10	151.066	315.941	490.653	554.818	271.445	327.275	149.143	83.875	156.920	36.666	13.649	8.507	1.458	5.590	2567
2010/11	106.178	280.582	228.857	304.885	296.254	138.686	301.285	60.997	141.323	97.412	37.006	0.000	4.019	0.000	1997
2011/12	704.863	977.323	434.876	313.742	272.140	239.320	154.581	175.088	84.582	92.435	89.376	17.638	6.808	4,989	3676
2012/13	178.500	781.083	631.421	166.627	126.961	142.044	110.084	97.000	74.340	69.473	43.376	38.450	7.458	0.773	2468
2013/14	15.919	314.865	218.715	344.981	151.631	132.767	120.756	118.377	89.555	74.602	48.695	44.637	31.096	11.598	1718
2014/15	152.422	90,269	330.084	260.919	259.079	187.905	111.955	91.629	37.855	76.680	30.366	10.619	22.799	10.108	1667
2015/16	381.900	164.221	174.507	312.350	225.836	215.207	93.743	62.753	75.339	41.961	15.696	26.756	20.159	5.401	1816
2016/17	97.036	220.642	137.217	151.937	262.488	136.801	241.382	61.220	55.869	62.805	11.435	20.135	13.733	0.313	1473
2017/18	32.749	22.947	95.097	171.664	201.944	319.933	209.174	255.348	75.813	34.505	83.460	54.903	25.370	28.115	1611
2018/19	306.295	137.402	67.933	201.362	101.946	110.810	167.397	163.804	73.346	30.040	29.950	38.499	9.138	7.271	1445

Table 11.1.1.2. Icelandic summers-spawning herring. Number of fish aged (number of scales) and number of samples taken in the annual acoustic surveys in the seasons 1987/88–2018/19 (age refers to the former year, i.e. autumns). In 2000 seven samples were used from the fishery. No survey was conducted in 1994/95.

	Number of scales															N of samples		
Year Age	2	3	4	5	6	7	8	9	10	11	12	13	14	15+	Total	Total	West	East
1987/88	11	59	246	156	37	28	58	33	22	16	23	10	5	8	712	8	1	7
1988/89	229	78	181	424	178	69	50	77	42	29	23	13	7	12	1412	18	5	10
1989/90	38	245	96	132	225	35	2	2	3	3	2	0	0	0	783	8		8
1990/91	418	229	303	90	131	257	28	6	3	8	0	0	0	0	1473	15		15
1991/92	414	439	127	127	33	48	84	5	3	0	2	0	0	1	1283	15		15
1992/93	122	513	289	68	73	28	38	34	6	2	2	6	0	0	1181	12		12
1993/94	63	285	343	129	13	15	7	14	11	0	1	3	0	0	884	9		9
1994/95*																		
1995/96	183	90	471	162	209	107	38	18	8	14	18	2	0	0	1320	14	9	5
1996/97	24	150	88	351	141	137	87	32	15	10	7	14	4	2	1062	11	4	7
1997/98	101	249	50	36	159	95	122	62	21	13	8	15	8	5	944	14	7	7
1998/99	130	216	777	72	31	65	59	86	37	22	17	5	6	11	1534	17	10	7
1999/00	116	227	72	144	17	13	26	26	27	10	8	2	1	0	689	7	3	4
2000/01	116	249	332	87	166	10	7	21	8	14	11	3	1	0	1025	14	10	4
2001/02	61	56	130	114	62	136	25	24	17	21	17	10	3	0	676	9	4	5
2002/03	520	705	258	104	130	74	128	46	26	25	13	15	10	1	2055	22	12	10
2003/04	126	301	415	88	35	32	15	17	3	4	4	6	1	1	1048	13	8	5
2004/05	304	159	284	326	70	29	17	5	8	4	0	3	3	0	1212	13	4	9
2005/06	217	312	190	420	501	110	40	38	26	18	5	5	5	7	1894	22	14	8
2006/07	19	77	134	64	71	88	22	4	2	2	0	0	0	1	484	6	4	2
2007/08	58	288	180	264	85	80	104	19	15	2	2	6	1	3	1107	17	13	4
2008/09	274	208	213	136	204	123	125	97	18	13	9	7	4	17	1448	29	19	10
2009/10	104	100	105	116	60	74	34	19	36	8	3	4	2	2	667	17	10	7
2010/11	35	74	102	157	139	61	119	22	52	36	13	0	1	0	811	11	8	3
2011/12	229	330	134	115	100	106	74	87	45	48	51	10	3	3	1335	15	9	6
2012/13‡	42	266	554	273	220	252	198	165	126	114	69	61	12	2	2370	60	55‡	5

	Number of scales															N of samples		
Year Age	2	3	4	5	6	7	8	9	10	11	12	13	14	15+	Total	Total	West	East
2013/14	26	472	275	414	199	200	199	208	163	138	90	85	60	23	2552	45	37‡	8
2014/15	83	50	96	71	72	53	32	26	11	22	8	3	6	4	534	10	8	2
2015/16	229	112	131	208	148	123	47	32	32	22	13	7	12	4	1120	14	7	7§
2016/17	66	164	122	137	202	117	169	43	50	44	14	15	9	4	1162	14	12	2
2017/18	35	58	82	77	75	101	65	77	29	11	27	18	8	9	672	10	5	5
2018/19	28	39	31	98	50	53	77	75	36	15	15	21	5	4	547	7	5	2

***No survey**

‡Samples in the western part were mainly from the commercial catch as there was impossible to secure a usable research survey samples from Kolgrafafjörður where most of the herring was observed.

§Three samples were taken in the east and south in this survey (B1-2016), while four were taken in the west and used also in the age-length key.

Table 11.2.1. Icelandic summer spawners. Landings, catches, recommended TACs, and set National TACs in thousand tonnes.

Year	Landings	Catches	Recom. TACs	Nat. TACs	Year	Landings	Catches	Recom. TACs	Nat. TACs
1972	0.31	0.31			2007/2008	158.9	158.9	130	150
1973	0.254	0.254			2008/2009	151.8	151.8	130	150
1974	1.275	1.275			2009/2010	46.3	46.3	40	47
1975	13.28	13.28			2010/2011	43.5	43.5	40	40
1976	17.168	17.168			2011/2012‡	49.4	49.4	40	45
1977	28.925	28.925			2012/2013‡	72.0	72.0	67	68.5
1978	37.333	37.333			2013/2014‡	72.0	72.0	87	87
1979	45.072	45.072			2014/2015‡§	95.0	95.0	83	83
1980	53.268	53.268			2015/2016‡	69.7	69.7	71	71
1981	39.544	39.544			2016/2017‡	60.4	60.4	63	63
1982	56.528	56.528			2017/2018‡	35.0	35.0	39	39
1983	58.867	58.867			2018/2019‡	40.7	40.7	35.1	35.1
1984	50.304	50.304							
1985	49.368	49.368	50	50					
1986	65.5	65.5	65	65					
1987	75	75	70	73					
1988	92.8	92.8	90	90					
1989	97.3	101	90	90					
1990/1991	101.6	105.1	80	110					
1991/1992	98.5	109.5	80	110					
1992/1993	106.7	108.5	90	110					
1993/1994	101.5	102.7	90	100					
1994/1995	132	134	120	120					
1995/1996	125	125.9	110	110					
1996/1997	95.9	95.9	100	100					
1997/1998	64.7	64.7	100	100					
1998/1999**	87	87	90	70					
1999/2000	92.9	92.9	100	100					
2000/2001	100.3	100.3	110	110					
2001/2002	95.7	95.7	125	125					
2002/2003*	96.1	96.1	105	105					
2003/2004*	130.7	130.7	110	110					
2004/2005	114.2	114.2	110	110					
2005/2006	103	103	110	110					
2006/2007	135	135	130	130					

***Summer fishery in 2002 and 2003 included**

**** TAC was decided 70 thousand tonnes but because of transfers from the previous quota year the national TAC became 90 thousand tonnes.**

‡ Landings and catches include bycatch of Icelandic summer-spawning herring in the mackerel and NSS herring fishery during the preceding summer (i.e. from the fishing season before in June–August).

§ The landings and catches in 2014/2015 consist of transfer of 7 kt from the year before and 5 kt from the year to come, which explains the discrepancy to the TACs.

Table 11.2.2.1. Icelandic summer-spawning herring. Catch in numbers (millions) and total catch in weight (thousand tonnes) (1981 refers to season 1981/1982 etc).

Year\ age	2	3	4	5	6	7	8	9	10	11	12	13	14	15+	Catch
1975	1.518	2.049	31.975	6.493	7.905	0.863	0.442	0.345	0.114	0.004	0.001	0.001	0.001	0.001	13.280
1976	0.614	9.848	3.908	34.144	7.009	5.481	1.045	0.438	0.296	0.134	0.092	0.001	0.001	0.001	17.168
1977	0.705	18.853	24.152	10.404	46.357	6.735	5.421	1.395	0.524	0.362	0.027	0.128	0.001	0.001	28.925
1978	2.634	22.551	50.995	13.846	8.738	39.492	7.253	6.354	1.616	0.926	0.4	0.017	0.025	0.051	37.333
1979	0.929	15.098	47.561	69.735	16.451	8.003	26.04	3.05	1.869	0.494	0.439	0.032	0.054	0.006	45.072
1980	3.147	14.347	20.761	60.727	65.328	11.541	9.285	19.442	1.796	1.464	0.698	0.001	0.11	0.079	53.268
1981	2.283	4.629	16.771	12.126	36.871	41.917	7.299	4.863	13.416	1.032	0.884	0.760	0.101	0.062	39.544
1982	0.454	19.187	28.109	38.280	16.623	38.308	43.770	6.813	6.633	10.457	2.354	0.594	0.075	0.211	56.528
1983	1.475	22.499	151.718	30.285	21.599	8.667	14.065	13.713	3.728	2.381	3.436	0.554	0.100	0.003	58.867
1984	0.421	18.015	32.244	141.354	17.043	7.113	3.916	4.113	4.517	1.828	0.202	0.255	0.260	0.003	50.304
1985	0.112	12.872	24.659	21.656	85.210	11.903	5.740	2.336	4.363	4.053	2.773	0.975	0.480	0.581	49.368
1986	0.100	8.172	33.938	23.452	20.681	77.629	18.252	10.986	8.594	9.675	7.183	3.682	2.918	1.788	65.500
1987	0.029	3.144	44.590	60.285	20.622	19.751	46.240	15.232	13.963	10.179	13.216	6.224	4.723	2.280	75.439
1988	0.879	4.757	41.331	99.366	69.331	22.955	20.131	32.201	12.349	10.250	7.378	7.284	4.807	1.957	92.828
1989	3.974	22.628	26.649	77.824	188.654	43.114	8.116	5.897	7.292	4.780	3.449	1.410	0.844	0.348	101.000
1990	12.567	14.884	56.995	35.593	79.757	157.225	30.248	8.187	4.372	3.379	1.786	0.715	0.446	0.565	105.097
1991	37.085	88.683	49.081	86.292	34.793	55.228	110.132	10.079	4.155	2.735	2.003	0.519	0.339	0.416	109.489
1992	16.144	94.86	122.626	38.381	58.605	27.921	38.42	53.114	11.592	1.727	1.757	0.153	0.376	0.001	108.504
1993	2.467	51.153	177.78	92.68	20.791	28.56	13.313	19.617	15.266	4.254	0.797	0.254	0.001	0.001	102.741
1994	5.738	134.616	113.29	142.876	87.207	24.913	20.303	16.301	15.695	14.68	2.936	1.435	0.244	0.195	134.003
1995	4.555	20.991	137.232	86.864	109.14	76.78	21.361	15.225	8.541	9.617	7.034	2.291	0.621	0.235	125.851
1996	0.717	15.969	40.311	86.187	68.927	84.66	39.664	14.746	8.419	5.836	3.152	5.18	1.996	0.574	95.882
1997	2.008	39.24	30.141	26.307	36.738	33.705	31.022	22.277	8.531	3.383	1.141	10.296	0.947	2.524	64.682
1998	23.655	45.39	175.529	22.691	8.613	40.898	25.944	32.046	14.647	2.122	2.754	2.15	1.07	1.011	86.998
1999	5.306	56.315	54.779	140.913	16.093	13.506	31.467	19.845	22.031	12.609	2.673	2.746	1.416	2.514	92.896
2000	17.286	57.282	136.278	49.289	76.614	11.546	8.294	16.367	9.874	11.332	6.744	2.975	1.539	1.104	100.332
2001	27.486	42.304	86.422	93.597	30.336	54.491	10.375	8.762	12.244	9.907	8.259	6.088	1.491	1.259	95.675
2002	11.698	80.863	70.801	45.607	54.202	21.211	42.199	9.888	4.707	6.52	9.108	9.355	3.994	5.697	96.128
2003	24.477	211.495	286.017	58.120	27.979	25.592	14.203	10.944	2.230	3.424	4.225	2.562	1.575	1.370	130.741
2004	23.144	63.355	139.543	182.45	40.489	13.727	9.342	5.769	7.021	3.136	1.861	3.871	0.994	1.855	114.237
2005	6.088	26.091	42.116	117.91	133.437	27.565	12.074	9.203	5.172	5.116	1.045	1.706	2.11	0.757	103.043
2006	52.567	118.526	217.672	54.800	48.312	57.241	13.603	5.994	4.299	0.898	1.626	1.213	0.849	0.933	135.303
2007	10.817	94.250	83.631	163.294	61.207	87.541	92.126	23.238	11.728	7.319	2.593	4.961	2.302	1.420	158.917
2008	10.427	38.830	90.932	79.745	107.644	59.656	62.194	54.345	18.130	8.240	5.157	2.680	2.630	1.178	151.780
2009	5.431	21.856	35.221	31.914	18.826	22.725	10.425	9.213	9.549	2.238	1.033	0.768	0.406	0.298	46.332
2010	1.476	8.843	22.674	29.492	24.293	14.419	17.407	10.045	7.576	8.896	1.764	1.105	0.672	0.555	43.533
2011	0.521	9.357	24.621	20.046	22.869	23.706	13.749	16.967	10.039	7.623	7.745	1.441	0.618	0.785	49.446
2012*	0.403	17.827	89.432	51.257	43.079	51.224	41.846	34.653	27.215	24.946	15.473	13.575	2.595	0.253	125.369
2013	6.888	46.848	24.833	35.070	17.250	18.550	19.032	21.821	15.952	15.804	10.081	9.775	6.722	2.486	72.058
2014	0.000	3.537	53.241	50.609	70.044	34.393	22.084	22.138	13.298	17.761	7.974	4.461	2.862	1.746	94.975
2015	0.089	6.024	29.89	53.573	43.501	43.015	15.533	10.76	8.664	8.161	6.981	2.726	2.467	1.587	69.729
2016	0.072	10.740	25.575	29.908	41.952	25.823	24.925	9.516	7.734	6.088	4.284	7.154	3.108	0.827	60.403
2017	1.262	5.236	31.855	18.113	10.239	15.506	10.223	8.830	5.676	3.399	1.616	2.220	1.533	1.596	35.034
2018	0.000	8.911	19.642	34.284	16.847	12.376	17.161	6.978	7.379	3.482	1.713	1.153	2.159	0.489	40.683

*** Includes both the landings (73.4 kt) and the herring that died in the mass mortality (52.0 kt) in the winter 2012/13 in Kolgrafafjörður.**

Table 11.2.2.2. Icelandic summer-spawning herring. The mean weight (g) at age from the commercial catch (1981 refers to season 1981/1982 etc.).

Year\age	2	3	4	5	6	7	8	9	10	11	12	13	14	15+
1975	110	179	241	291	319	339	365	364	407	389	430	416	416	416
1976	103	189	243	281	305	335	351	355	395	363	396	396	396	396
1977	84	157	217	261	285	313	326	347	364	362	358	355	400	420
1978	73	128	196	247	295	314	339	359	360	376	380	425	425	425
1979	75	145	182	231	285	316	334	350	367	368	371	350	350	450
1980	69	115	202	232	269	317	352	360	380	383	393	390	390	390
1981	61	141	190	246	269	298	330	356	368	405	382	400	400	400
1982	65	141	186	217	274	293	323	354	385	389	400	394	390	420
1983	59	132	180	218	260	309	329	356	370	407	437	459	430	472
1984	49	131	189	217	245	277	315	322	351	334	362	446	417	392
1985	53	146	219	266	285	315	335	365	388	400	453	469	433	447
1986	60	140	200	252	282	298	320	334	373	380	394	408	405	439
1987	60	168	200	240	278	304	325	339	356	378	400	404	424	430
1988	75	157	221	239	271	298	319	334	354	352	371	390	408	437
1989	63	130	206	246	261	290	331	338	352	369	389	380	434	409
1990	80	127	197	245	272	285	305	324	336	362	370	382	375	378
1991	74	135	188	232	267	289	304	323	340	352	369	402	406	388
1992	68	148	190	235	273	312	329	339	355	382	405	377	398	398
1993	66	145	211	246	292	324	350	362	376	386	419	389	389	389
1994	66	134	201	247	272	303	333	366	378	389	390	412	418	383
1995	68	130	183	240	277	298	325	358	378	397	409	431	430	467
1996	75	139	168	212	258	289	308	325	353	353	377	404	395	410
1997	63	131	191	233	269	300	324	341	355	362	367	393	398	411
1998	52	134	185	238	264	288	324	340	348	375	406	391	426	456
1999	74	137	204	233	268	294	311	339	353	362	378	385	411	422
2000	62	159	217	268	289	325	342	363	378	393	407	425	436	430
2001	74	139	214	244	286	296	324	347	354	385	403	421	421	433
2002	85	161	211	258	280	319	332	354	405	396	416	433	463	460
2003	72	156	189	229	260	283	309	336	336	369	394	378	412	423
2004	84	149	213	248	280	315	331	349	355	379	388	412	419	425
2005	106	170	224	262	275	298	324	335	335	356	372	394	405	413
2006	107	189	234	263	290	304	339	349	369	416	402	413	413	467
2007	93	158	221	245	261	277	287	311	339	334	346	356	384	390
2008	105	174	232	275	292	307	315	327	345	366	377	372	403	434
2009	113	190	237	274	304	318	326	335	342	360	372	394	409	421
2010	87	204	243	271	297	315	329	335	341	351	367	366	405	416
2011	97	187	245	283	309	328	343	352	356	364	375	386	378	432
2012	65	206	244	282	301	320	333	344	350	359	364	367	373	391
2013	95	182	238	271	300	322	337	349	360	365	362	375	377	394
2014		202	259	288	306	328	346	354	362	366	367	380	383	403
2015	107	203	249	275	299	313	329	347	352	358	361	368	380	378
2016	129	202	242	281	303	322	336	355	359	368	369	379	386	402
2017	95	192	252	281	303	324	341	350	367	376	384	389	395	402
2018		191	252	293	317	333	347	350	366	375	389	388	392	383

Table 11.2.2.3. Icelandic summer-spawning herring. Proportion mature at age (1981 refers to season 1981/1982 etc.).

Year\age	2	3	4	5	6	7	8	9	10	11	12	13	14	15+
1975	0	0.27	0.97	1	1	1	1	1	1	1	1	1	1	1
1976	0	0.13	0.9	1	1	1	1	1	1	1	1	1	1	1
1977	0	0.02	0.87	1	1	1	1	1	1	1	1	1	1	1
1978	0	0.04	0.78	1	1	1	1	1	1	1	1	1	1	1
1979	0	0.07	0.65	0.98	1	1	1	1	1	1	1	1	1	1
1980	0	0.05	0.92	1	1	1	1	1	1	1	1	1	1	1
1981	0	0.03	0.65	0.99	1	1	1	1	1	1	1	1	1	1
1982	0.02	0.05	0.85	1	1	1	1	1	1	1	1	1	1	1
1983	0	0	0.64	1	1	1	1	1	1	1	1	1	1	1
1984	0	0.01	0.82	1	1	1	1	1	1	1	1	1	1	1
1985	0	0	0.9	1	1	1	1	1	1	1	1	1	1	1
1986–2018	0	0.2	0.85	1	1	1	1	1	1	1	1	1	1	1

Table 11.3.2.1. Icelandic summer-spawning herring. Natural mortality at age for the different years (refers to the autumn) where the deviation from the fixed M = 0.1 is due to the *Ichthyophonus* infection (1981 refers to season 1981/1982 etc.). The estimate of, for example, M for age 4 in 2018 represents estimated infection rate of age 3 in 2017.

Year\age	3	4	5	6	7	8	9	10	11	12	13	14	15	13+
1987–2008	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10
2009*	0.22	0.22	0.22	0.22	0.22	0.22	0.22	0.22	0.22	0.22	0.22	0.22	0.22	0.22
2010*	0.29	0.29	0.28	0.26	0.25	0.24	0.24	0.24	0.23	0.23	0.23	0.23	0.23	0.23
2011*	0.13	0.26	0.26	0.25	0.23	0.24	0.25	0.24	0.20	0.21	0.21	0.21	0.21	0.21
2012-2016	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10	0.10
2017**	0.111	0.118	0.124	0.173	0.175	0.175	0.207	0.187	0.256	0.279	0.210	0.180	0.191	0.183
2018**	0.116	0.112	0.172	0.162	0.175	0.228	0.226	0.247	0.275	0.338	0.307	0.184	0.186	0.250
2019***	0.111	0.135	0.144	0.168	0.216	0.169	0.171	0.183	0.245	0.189	0.243	0.182	0.140	0.189

*** Based on prevalence of infection estimates and acoustic measurements ($M_{infected}$ multiplied by 0.3 and added to 0.1; Óskarsson *et al.* 2018b).**

**** Based on prevalence of infection estimates in the winter 2016/17 and 2017/18 (multiplied by 0.3 and added to 0.1; Óskarsson and Pálsson, 2017; 2018).**

***** Based on prevalence of infection estimates in the winter 2018/19 (multiplied by 0.3 and added to 0.1) and should by applied in the prognosis in the 2019 assessment.**

Table 11.3.2.2. Model settings and results of model parameters from the final NFT-Adapt run in 2019 for Icelandic summer spawning herring.

VPA Version 3.3.0

Model ID: Corrected run 2019

Input File: C:\HAFRONET_GOGN\NWWG OG UTTEKTIR\NWWG2019\RUN4\RUN4.DAT

Date of Run: 10-APR-2019 Time of Run: 09:23

Levenburg-Marquardt Algorithm Completed 6 Iterations

Residual Sum of Squares = 54.3004

Number of Residuals = 248

Number of Parameters = 9

Degrees of Freedom = 239

Mean Squared Residual = 0.227198

Standard Deviation = 0.476653

Number of Years = 32

Number of Ages = 11

First Year = 1987

Youngest Age = 3

Oldest True Age = 12

Number of Survey Indices Available = 10

Number of Survey Indices Used in Estimate = 8

VPA Classic Method - Auto Estimated Q's

Stock Numbers Predicted in Terminal Year Plus One (2019)

Age	Stock Predicted	Std. Error	CV
4	133794.780	0.648712E+05	0.484856E+00
5	29940.160	0.123475E+05	0.412404E+00
6	91689.598	0.319492E+05	0.348449E+00
7	69140.315	0.212237E+05	0.306966E+00
8	61111.490	0.173486E+05	0.283884E+00
9	94641.491	0.250973E+05	0.265183E+00
10	71523.309	0.178582E+05	0.249684E+00
11	63403.473	0.159617E+05	0.251748E+00
12	21024.211	0.656008E+04	0.312025E+00

Catchability Values for Each Survey Used in Estimate

INDEX Catchability Std. Error CV

1	0.102696E+01	0.923730E-01	0.899479E-01
2	0.130889E+01	0.113718E+00	0.868813E-01
3	0.137106E+01	0.856488E-01	0.624693E-01
4	0.144946E+01	0.919799E-01	0.634578E-01
5	0.155782E+01	0.118033E+00	0.757683E-01
6	0.177129E+01	0.145467E+00	0.821245E-01
7	0.186284E+01	0.198632E+00	0.106628E+00
8	0.176619E+01	0.197296E+00	0.111707E+00

-- Non-Linear Least Squares Fit --

Maximum Marquadt Iterations = 100

Scaled Gradient Tolerance = 6.055454E-05

Scaled Step Tolerance = 1.000000E-18

Relative Function Tolerance = 1.000000E-18

Absolute Function Tolerance = 4.930381E-32

Reported Machine Precision = 2.220446E-16

VPA Method Options

- Catchability Values Estimated as an Analytic Function of N
- Catch Equation Used in Cohort Solution
- Plus Group Forward Calculation Method Used
- Arithmetic Average Used in F-Oldest Calculation
- F-Oldest Calculation in Years Prior to Terminal Year
  Uses Fishing Mortality in Ages 8 to 11
- Calculation of Population of Age 3 In Year 2019
  = Geometric Mean of First Age Populations
  Year Range Applied = 1991 to 2013
- Survey Weight Factors Were Used

Stock Estimates for Age 4 to Age 12

Full F in Terminal Year = 0.1613

F in Oldest True Age in Terminal Year = 0.1160

Full F Calculated Using Classic Method

F in Oldest True Age in Terminal Year has been

Calculated in Same Manner as in All Other Years

Age Input Partial Calc Partial Fishing Used In

Recruitment Recruitment Mortality Full F Comments

3	0.500	0.127	0.0609	NO	Stock Estimate in T+1
4	0.800	1.000	0.4807	NO	Stock Estimate in T+1
5	1.000	0.611	0.2936	YES	Stock Estimate in T+1
6	1.000	0.420	0.2020	YES	Stock Estimate in T+1
7	1.000	0.353	0.1696	YES	Stock Estimate in T+1
8	1.000	0.311	0.1493	YES	Stock Estimate in T+1
9	1.000	0.173	0.0833	YES	Stock Estimate in T+1
10	1.000	0.203	0.0975	YES	Stock Estimate in T+1
11	1.000	0.279	0.1340	YES	Stock Estimate in T+1
12	1.000	0.241	0.1160		F-Oldest

Table 11.3.2.3. Icelandic summer spawners stock estimates (from NFT-Adapt in 2019) in numbers (millions) by age (years) at 1 January during 1987–2019.

Year\Age	3	4	5	6	7	8	9	10	11	12	13+	Total
1987	529.83	988.97	300.67	84.60	69.14	107.46	42.63	38.03	26.41	34.26	34.29	2256
1988	271.00	476.42	852.47	214.85	56.99	43.83	53.49	24.15	21.19	14.26	36.99	2066
1989	447.33	240.69	391.81	676.97	128.70	29.84	20.62	18.03	10.18	9.48	26.10	2000
1990	300.83	383.26	192.47	280.67	433.68	75.61	19.30	13.07	9.41	4.69	26.46	1739
1991	840.56	258.05	292.67	140.37	178.35	243.51	39.78	9.72	7.68	5.31	24.86	2041
1992	1033.12	676.33	186.91	183.02	94.01	109.04	116.17	26.44	4.86	4.36	24.19	2458
1993	635.46	844.69	495.58	132.71	110.07	58.60	62.27	54.88	12.96	2.77	23.67	2434
1994	691.75	526.39	595.62	360.46	100.34	72.51	40.39	37.75	35.19	7.69	22.92	2491
1995	202.72	498.17	368.80	403.41	243.44	67.16	46.36	21.12	19.31	17.95	23.14	1912
1996	181.40	163.49	320.64	251.31	261.54	147.51	40.53	27.52	11.03	8.38	27.53	1441
1997	772.62	148.97	109.70	208.41	162.04	156.43	95.86	22.71	16.92	4.46	22.16	1720
1998	320.53	661.80	106.19	74.31	153.70	114.64	112.10	65.61	12.47	12.10	10.03	1643
1999	552.74	246.93	432.37	74.56	59.06	100.29	79.12	71.05	45.47	9.27	13.41	1684
2000	391.54	446.64	171.46	257.71	52.19	40.62	60.93	52.76	43.41	29.19	11.67	1558
2001	469.04	299.89	274.97	108.41	160.56	36.27	28.89	39.61	38.37	28.53	25.26	1510
2002	1457.99	384.21	189.43	160.14	69.34	93.66	22.98	17.83	24.24	25.33	32.47	2478
2003	1077.32	1242.40	280.45	128.14	93.55	42.63	44.83	11.44	11.67	15.75	25.70	2974
2004	666.57	774.09	852.84	198.61	89.40	60.38	25.12	30.19	8.24	7.32	28.28	2741
2005	994.34	542.95	567.97	598.57	141.29	67.86	45.76	17.26	20.65	4.48	24.06	3025
2006	739.70	874.92	451.26	402.04	415.01	101.69	49.94	32.68	10.71	13.84	20.50	3112
2007	665.25	556.78	585.21	356.27	317.89	321.16	79.09	39.50	25.48	8.84	26.68	2982
2008	532.42	512.97	425.74	377.84	262.03	202.70	202.02	49.36	24.59	16.09	21.45	2627
2009	454.85	444.86	377.84	309.54	239.83	180.50	124.46	131.27	27.50	14.44	22.93	2328
2010	442.79	346.57	326.59	275.61	232.33	172.74	135.97	91.95	97.12	20.13	27.84	2170
2011	604.09	323.05	239.32	222.04	191.49	168.26	120.01	97.98	65.77	69.13	34.45	2136
2012	424.15	521.68	227.81	167.72	152.85	130.84	120.47	78.84	68.36	46.88	74.86	2014
2013	509.46	366.84	387.15	157.50	110.90	89.77	78.73	76.16	45.55	38.23	79.91	1940
2014	230.15	416.47	308.33	316.99	126.13	82.74	63.17	50.55	53.77	26.25	79.33	1754
2015	197.04	204.89	326.27	230.95	220.37	81.52	53.92	36.19	33.13	31.83	79.36	1495
2016	230.17	172.57	157.01	244.37	167.68	158.57	59.02	38.58	24.53	22.24	87.54	1362
2017	66.04	198.06	131.86	113.68	181.29	127.21	119.82	44.37	27.57	16.42	84.74	1111
2018	159.68	54.16	146.06	99.50	86.26	138.02	97.44	89.48	31.65	18.37	76.04	997
2019	360.00	133.80	29.94	91.69	69.14	61.11	94.64	71.52	63.40	21.02	67.55	1064

*** Number at age 3 in 2019 is predicted from a survey index of number at age 1 in 2017 (see Section 11.6.1).**

Table 11.3.2.4. Estimated fishing mortality at age of Icelandic summer-spawning herring (from NFT-Adapt in 2019) by age (years) during 1987–2018 (referring to the autumn of the fishing season) and weighed average F by numbers for age 5–10.

Year\Age	3	4	5	6	7	8	9	10	11	12	13+	WF5-10
1987	0.006	0.049	0.236	0.295	0.356	0.598	0.468	0.485	0.516	0.517	0.517	0.347
1988	0.019	0.096	0.131	0.412	0.547	0.654	0.988	0.764	0.704	0.777	0.506	0.266
1989	0.055	0.124	0.234	0.345	0.432	0.336	0.356	0.550	0.674	0.479	0.111	0.322
1990	0.053	0.170	0.216	0.353	0.477	0.542	0.586	0.431	0.472	0.508	0.071	0.400
1991	0.117	0.223	0.369	0.301	0.392	0.640	0.309	0.592	0.466	0.502	0.055	0.436
1992	0.101	0.211	0.243	0.409	0.373	0.460	0.650	0.613	0.465	0.547	0.023	0.415
1993	0.088	0.249	0.218	0.180	0.317	0.272	0.400	0.345	0.421	0.360	0.011	0.248
1994	0.228	0.256	0.290	0.293	0.302	0.347	0.548	0.571	0.573	0.510	0.090	0.312
1995	0.115	0.341	0.284	0.333	0.401	0.405	0.421	0.550	0.735	0.528	0.154	0.343
1996	0.097	0.299	0.331	0.339	0.414	0.331	0.479	0.386	0.804	0.500	0.350	0.361
1997	0.055	0.239	0.290	0.205	0.246	0.233	0.279	0.500	0.235	0.312	1.042	0.250
1998	0.161	0.326	0.254	0.130	0.327	0.271	0.356	0.267	0.197	0.273	0.582	0.280
1999	0.113	0.265	0.418	0.257	0.274	0.398	0.305	0.393	0.343	0.360	0.734	0.377
2000	0.167	0.385	0.358	0.373	0.264	0.241	0.331	0.219	0.320	0.277	0.699	0.335
2001	0.100	0.359	0.441	0.347	0.439	0.356	0.382	0.391	0.316	0.361	0.456	0.414
2002	0.060	0.215	0.291	0.438	0.386	0.637	0.598	0.324	0.331	0.472	0.946	0.418
2003	0.231	0.276	0.245	0.260	0.338	0.429	0.296	0.229	0.367	0.330	0.255	0.279
2004	0.105	0.210	0.254	0.241	0.176	0.177	0.276	0.280	0.508	0.310	0.287	0.244
2005	0.028	0.085	0.246	0.266	0.229	0.207	0.237	0.377	0.301	0.280	0.222	0.252
2006	0.184	0.302	0.136	0.135	0.156	0.151	0.135	0.149	0.092	0.132	0.166	0.143
2007	0.160	0.168	0.338	0.207	0.350	0.364	0.371	0.374	0.360	0.367	0.417	0.320
2008	0.080	0.206	0.219	0.355	0.273	0.388	0.331	0.485	0.432	0.409	0.381	0.308
2009	0.055	0.092	0.099	0.070	0.111	0.066	0.086	0.084	0.095	0.083	0.074	0.087
2010	0.023	0.078	0.109	0.105	0.073	0.120	0.087	0.097	0.108	0.103	0.098	0.099
2011	0.017	0.090	0.100	0.123	0.149	0.096	0.173	0.122	0.137	0.132	0.096	0.124
2012*	0.045	0.198	0.269	0.314	0.432	0.408	0.359	0.449	0.481	0.424	0.261	0.355
2013	0.102	0.074	0.100	0.122	0.193	0.251	0.343	0.248	0.451	0.324	0.286	0.164
2014	0.016	0.144	0.189	0.264	0.337	0.328	0.457	0.323	0.425	0.383	0.128	0.271
2015	0.033	0.166	0.189	0.220	0.229	0.223	0.235	0.289	0.299	0.261	0.094	0.215
2016	0.050	0.169	0.223	0.199	0.176	0.180	0.185	0.236	0.301	0.226	0.143	0.196
2017	0.087	0.187	0.158	0.103	0.098	0.092	0.085	0.151	0.150	0.119	0.072	0.110
2018	0.061	0.481	0.294	0.202	0.170	0.149	0.083	0.098	0.134	0.116	0.058	0.175

*** Derived from both the landings ($WF_{5–10}$ ~0.209) and the herring that died in the mass mortality (0.148) in the winter 2012/13 in Kolgrafafjörður (Óskarsson *et al.,* 2018a). WF5–10 without the mass mortality was 0.214.**

Table 11.3.2.5. Summary table from NFT-Adapt run in 2019 for Icelandic summer spawning herring.

Year	Recruits age 3 (millions)	Biomass age 3+ (kt)	Biomass age 4+ (kt)	SSB (kt)	Landings age 3+ (kt)	Yield/SSB	$WF_{age\ 5\text{-}10}$	HR 4+
1987	530	504	415	384	75	0.20	0.35	0.182
1988	271	495	452	423	93	0.22	0.27	0.205
1989	447	459	401	386	101	0.26	0.32	0.251
1990	301	410	371	350	104	0.30	0.40	0.281
1991	841	424	310	310	107	0.34	0.44	0.344
1992	1033	502	349	343	107	0.31	0.42	0.307
1993	635	546	454	424	103	0.24	0.25	0.226
1994	692	553	461	441	134	0.30	0.31	0.290
1995	203	462	435	406	125	0.31	0.34	0.288
1996	181	348	322	307	96	0.31	0.36	0.297
1997	773	368	267	269	65	0.24	0.25	0.243
1998	321	366	323	298	86	0.29	0.28	0.266
1999	553	373	297	290	93	0.32	0.38	0.312
2000	392	386	324	306	100	0.33	0.33	0.308
2001	469	348	283	272	94	0.34	0.41	0.331
2002	1458	513	278	298	96	0.32	0.42	0.345
2003	1077	580	412	390	129	0.33	0.28	0.313
2004	667	617	517	488	112	0.23	0.24	0.217
2005	994	708	539	527	102	0.19	0.25	0.190
2006	740	789	649	615	130	0.21	0.14	0.200
2007	665	703	598	572	158	0.28	0.32	0.264
2008	532	690	597	569	151	0.26	0.31	0.252
2009	455	636	549	494	46	0.09	0.09	0.083
2010	443	604	514	456	43	0.10	0.10	0.084
2011	604	590	477	433	49	0.11	0.12	0.104
2012*	424	560	473	448	125	0.28	0.36	0.265
2013	509	514	421	406	71	0.18	0.16	0.170
2014	230	511	464	435	95	0.22	0.27	0.205
2015	197	430	390	371	70	0.19	0.22	0.179
2016	230	399	353	338	60	0.18	0.20	0.171
2017	66	343	331	299	35	0.12	0.11	0.106
2018	160	311	280	256	41	0.16	0.18	0.145
2019	360[§]	301	230	228				
Mean	534	501	416	394	94	0.24	0.27	0.23

*** The mass mortality of 52 thousand tonnes in Kolgrafafjörður in the winter 2012/13 is not included in the landings, yield/SSB, or WF, even if included as landings in the analytical assessment.**

§ Number at age 3 in 2019 is predicted from a survey index of number at age 1 in 2017 (see Section 11.6.1).

Table 11.3.2.6. The residuals from survey observations and NFT-Adapt 2019 results for Icelandic summer spawning herring (no surveys in 1987 and 1995) on 1 January.

Year\Age	4	5	6	7	8	9	10	11
1987								
1988	-0.198	-0.263	0.045	-0.369	-0.741	-0.297	-0.203	-0.462
1989	-0.205	-0.790	-0.889	0.010	-0.001	-0.003	0.000	0.000
1990	0.510	-0.340	-0.322	-0.058	0.422	-0.433	-0.001	-0.002
1991	-0.695	-0.393	-0.713	-0.302	0.305	0.119	0.007	-0.004
1992	0.413	0.371	0.243	-0.417	-0.205	0.222	-0.839	0.001
1993	-0.043	0.118	-0.136	-0.199	-0.522	-0.135	-0.056	0.070
1994	-0.068	0.124	0.004	-0.776	-0.661	0.395	-0.364	-0.540
1995								
1996	-0.227	0.596	-0.215	0.016	-0.262	0.313	-0.055	-0.182
1997	0.571	-0.072	0.495	0.139	0.290	0.247	0.788	0.619
1998	-0.122	-0.539	-0.574	0.253	-0.135	0.024	-0.145	0.477
1999	0.009	0.648	0.011	-0.503	-0.144	-0.687	-0.264	-0.398
2000	0.604	0.063	0.539	0.154	-0.378	0.429	-0.089	0.459
2001	1.144	1.297	0.251	0.729	-0.497	-1.180	-0.665	-1.554
2002	-0.318	-0.131	0.171	0.472	0.863	0.428	0.542	-0.109
2003	0.408	0.410	0.158	0.660	0.835	1.247	1.538	0.837
2004	0.591	0.612	0.195	-0.172	0.069	-0.140	-0.210	-0.007
2005	0.247	0.319	0.243	-0.178	-0.527	-0.604	-1.077	-0.421
2006	-0.705	-0.538	0.396	0.708	0.575	0.323	0.756	1.355
2007	0.062	0.325	-0.175	-0.082	0.326	-0.378	0.520	0.080
2008	-0.143	-0.656	0.044	-0.205	0.244	0.675	0.880	1.729
2009	-0.846	-0.171	-0.387	0.275	-0.052	0.024	-0.370	-0.484
2010	-0.119	0.138	0.384	-0.216	0.196	-0.479	-0.714	-0.089
2011	-0.168	-0.314	0.001	0.065	-0.637	0.349	-1.096	0.196
2012	0.601	0.377	0.311	0.206	0.161	-0.322	0.176	-0.356
2013	0.729	0.220	-0.259	-0.236	0.015	-0.236	-0.380	-0.080
2014	-0.306	-0.613	-0.231	-0.187	0.030	0.076	0.229	-0.059
2015	-0.846	-0.258	-0.194	-0.209	0.392	0.159	0.307	-0.435
2016	-0.076	-0.164	-0.070	-0.073	-0.138	-0.109	-0.136	0.553
2017	0.081	-0.229	-0.026	-0.001	-0.371	0.129	-0.300	0.137
2018	-0.885	-0.698	0.230	0.479	0.397	0.192	0.427	0.305
2019	0.000	0.550	0.471	0.017	0.152	-0.001	0.207	-0.423
Max. Residuals	1.144	1.297	-0.889	-0.776	0.863	1.247	1.538	1.729

Table 11.6.1.1. The input data used for prognosis of the Icelandic summer-spawning herring in the 2019 assessment: the predicted weights, the selection pattern, M, proportion of M before spawning, and the number-at-age derived from NFT-Adapt run.

Age (year class)	Mean weights (kg)	M	Maturity ogive	Selection pattern	Mortality prop. before spawning		Number at age
					F	M	1 January 2019
3 (2016)	0.164	0.11	0.200	0.467	0.000	0.500	360.0
4 (2015)	0.239	0.14	0.850	0.863	0.000	0.500	133.8
5 (2014)	0.286	0.14	1.000	1.000	0.000	0.500	29.9
6 (2013)	0.318	0.17	1.000	1.000	0.000	0.500	91.7
7 (2012)	0.336	0.22	1.000	1.000	0.000	0.500	69.1
8 (2011)	0.349	0.17	1.000	1.000	0.000	0.500	61.1
9 (2010)	0.360	0.17	1.000	1.000	0.000	0.500	94.6
10 (2009)	0.362	0.18	1.000	1.000	0.000	0.500	71.5
11 (2008)	0.375	0.25	1.000	1.000	0.000	0.500	63.4
12 (2007)	0.382	0.19	1.000	1.000	0.000	0.500	21.0
13+ (2006+)	0.392	0.19	1.000	1.000	0.000	0.500	67.5

Table 11.6.2.1. Icelandic summer-spawning herring. Catch options table for the 2019/2020 season according to the Management plan where the basis is: SSB (1 July 2019) 212 kt (accounted for $M_{infection}$ in 2018); Biomass age 4+ (1 January 2019) is 231 kt; Catch (2018/19) 40 kt; HR (2018) 0.145, and WF_{5-10}(2018) 0.175. Other options are also shown, including MSY approach, where SSB_{2019} < MSY $B_{trigger}$ = 273 kt, hence resulting F is $F_{MSY} \times SSB_{2018}/B_{trigger} = 0.22 \times 212/273 = 0.171$.

Rationale	Catches (2019/2020)	Basis	F (2019/2020)	Biomass of age 4+ (2020)	SSB 2020	%SSB change *	% TAC change **
Management plan	**34.6**	**HR = 0.15**	**0.163**	**220**	**221**	**4**	**-2**
MSY approach	36.0	F_{MSY}	0.171		220	4	2
Zero catch	0	F = 0	0		251	18	-100
F_{pa}	45.4	F_{pa} = 0.22	0.220		212	0	29
F_{lim}	106.9	F_{lim} = 0.61	0.610		159	-25	204

***SSB 2020 relative to SSB 2019**

****TAC 2019/20 relative to landings 2018/19**

11.19 Figures

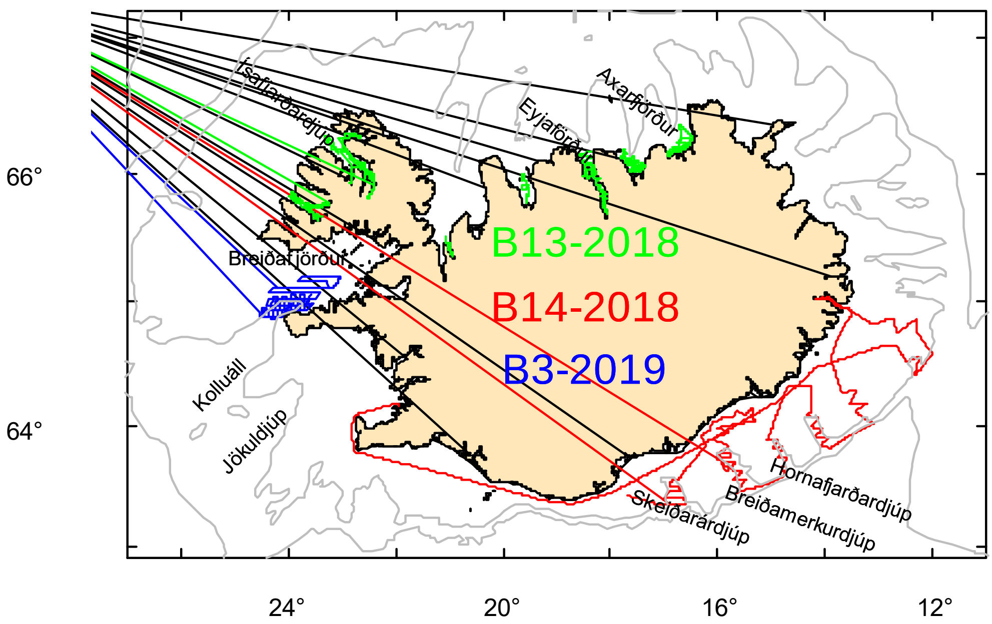

Figure 11.1.2.1. The survey tracks of three acoustic surveys on Icelandic summer-spawning herring in the north (B13-2018; juveniles), east and south (B3-2018; adult) and in the west (B3-2019; adult; blue) in 2018/19 and locations of the areas that are referred to in the text.

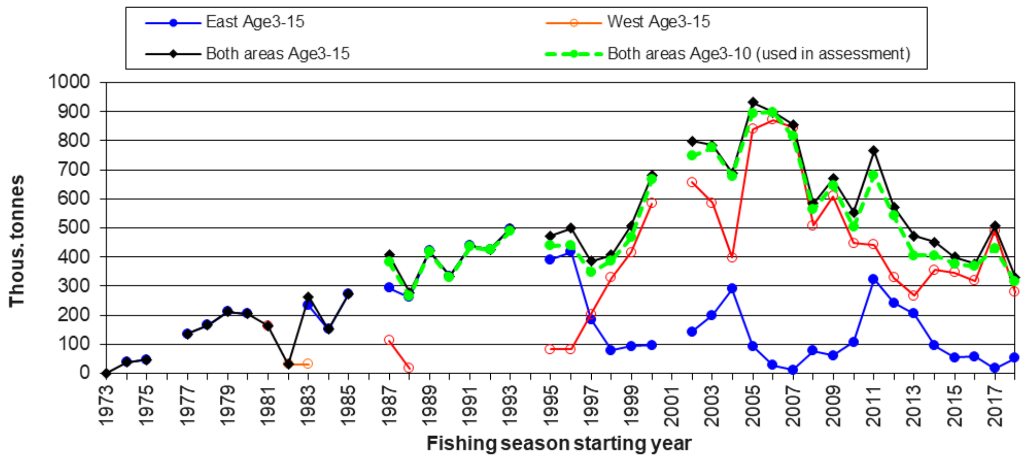

Figure 11.1.2.2 Total biomass index for Icelandic summer-spawning herring from the acoustic surveys for ages 3+ in the areas east and west of 18°W (except in 2011 and 2012 where fish outside of Breiðafjörður was set to the eastern part), combined over all areas and age 3–10 which are used in tuning of the analytical assessment. The years in the plot (1973–2018) refer to the autumn of the fishing seasons.

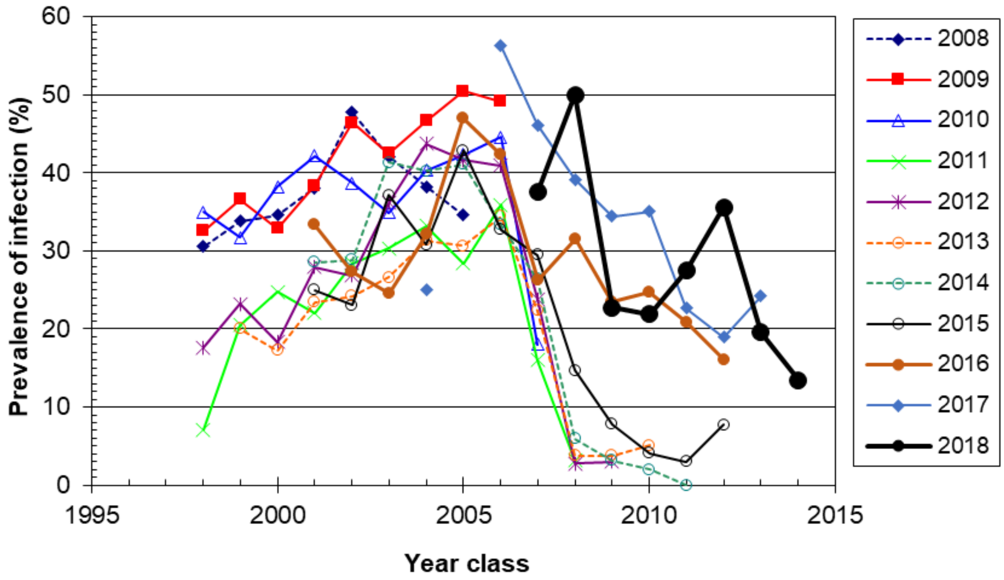

Figure 11.1.3.1. The prevalence of *Ichthyophonus* infection for the different year classes of Icelandic summer-spawning herring in Breiðafjörður (in 2008–2013) and west of Iceland (in 2014–2018) as estimated from catch samples in the autumns (Oct.-Dec.).The prevalence of *Ichthyophonus* infection for the different year classes of Icelandic summer-spawning herring in Breiðafjörður and west of Iceland as estimated from catch samples in the autumns 2008 to 2017.

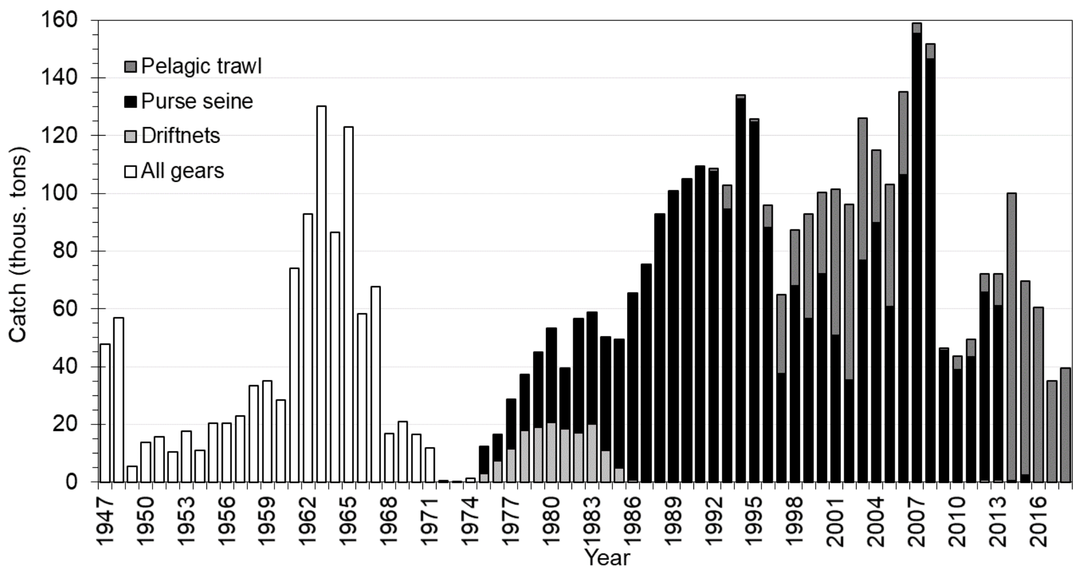

Figure 11.2.1. Icelandic summer spawning herring. Seasonal total landings (in thousand tonnes) during 1947–2018, referring to the autumns, by different fishing gears from 1975 onwards).

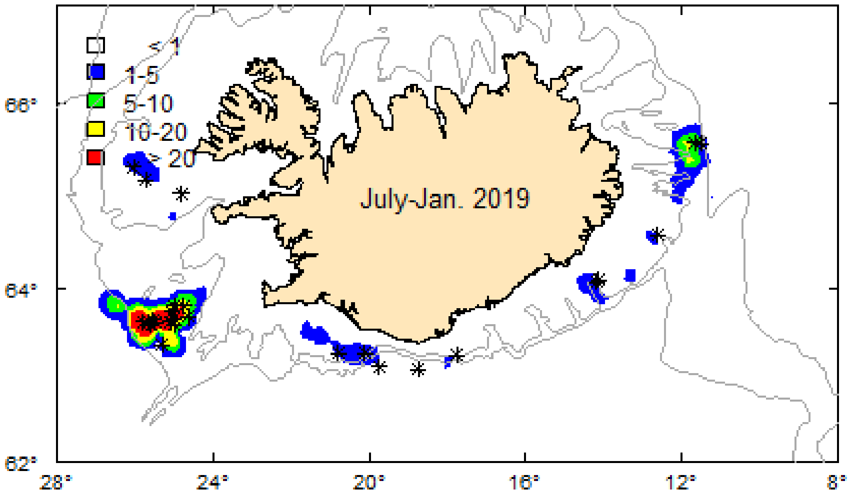

Figure 11.2.2. The distribution of the fishery (in tonnes) of Icelandic summer spawning herring during the fishing season 2018/19, including the bycatch in the mackerel fishery in July–September 2018. The stars indicate the location of catch samples.

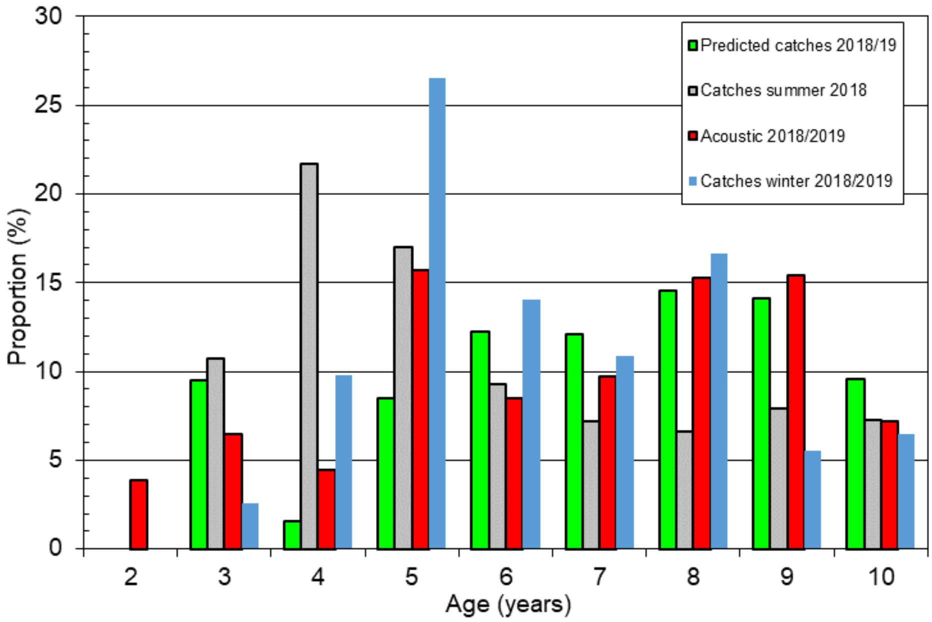

Figure 11.2.2.1. Proportion of the different age groups of Icelandic summer-spawning herring to the total catches (biomass) as observed in 2018/2019 fishing season (June 2018–January 2019), predicted catch composition the 2018/19 season from the 2018 assessment (ICES, 2018), and the summer catches in June–October 2018 in comparison to the age composition in the stock according to the acoustic measurements in the winter 2018/2019.

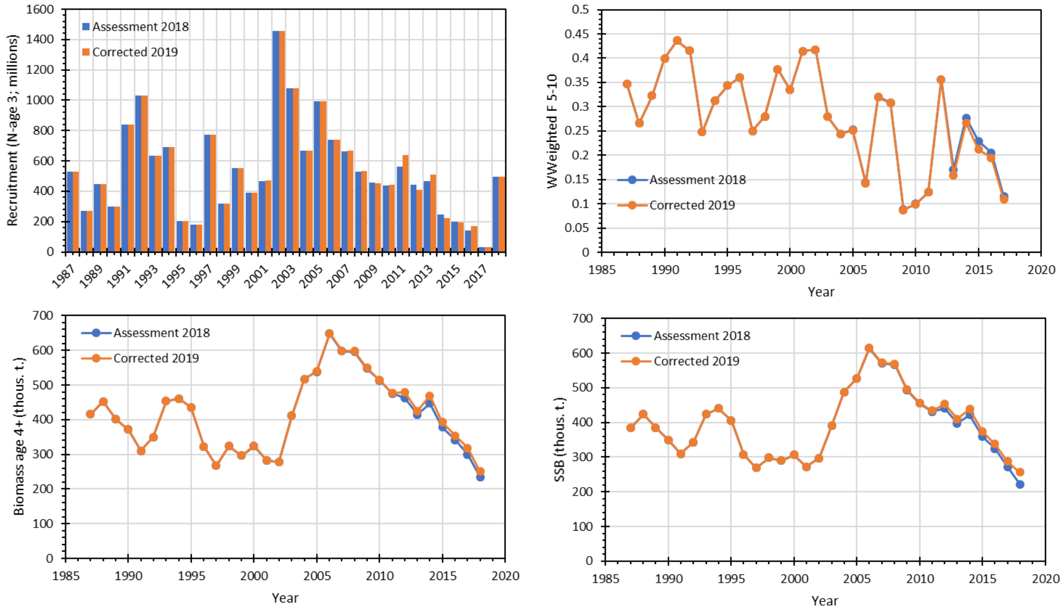

Figure 11.2.3.1. Comparison of the final assessment of Icelandic summer-spawning herring in 2018 (with error in the tuning series) and an assessment where the survey indices from 2017 have been corrected (i.e. a SPALY).

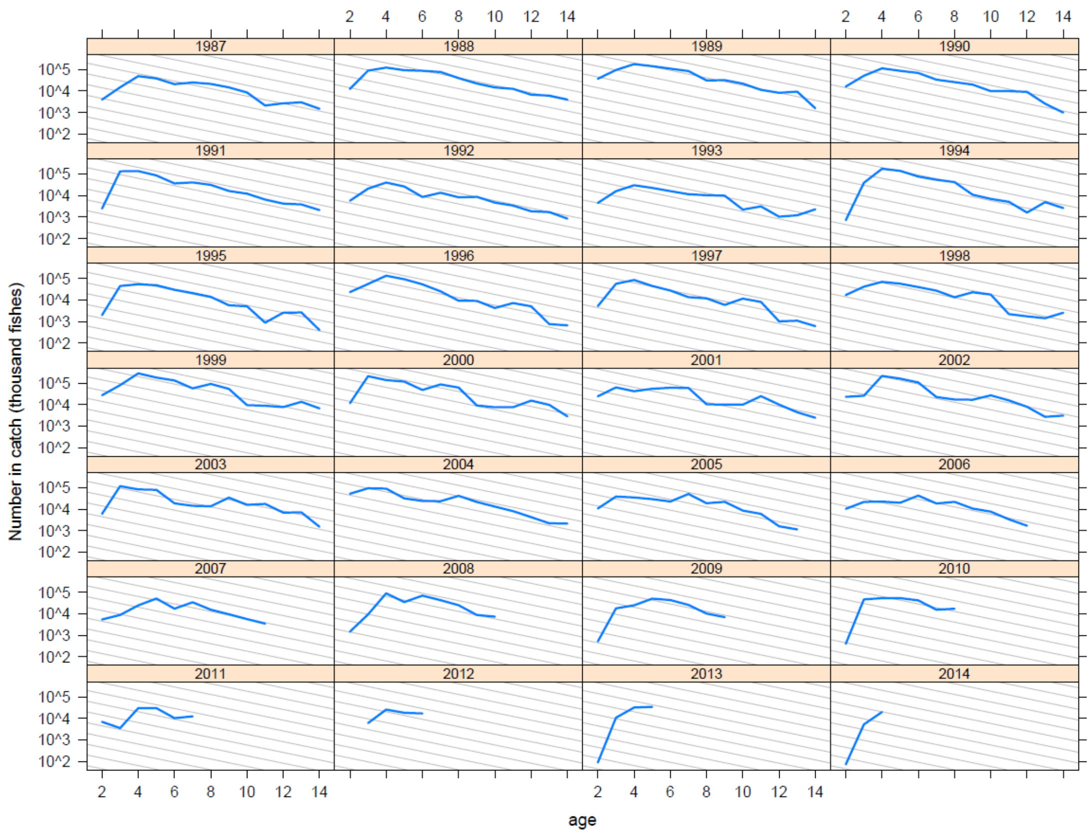

Figure 11.3.1.1. Icelandic summer-spawning herring. Catch curves (log_2 of catches) by year classes 1987–2014. Grey lines correspond to Z = 0.4. Note that the mass mortality in Kolgrafafjörður is added to the catches in 2012.

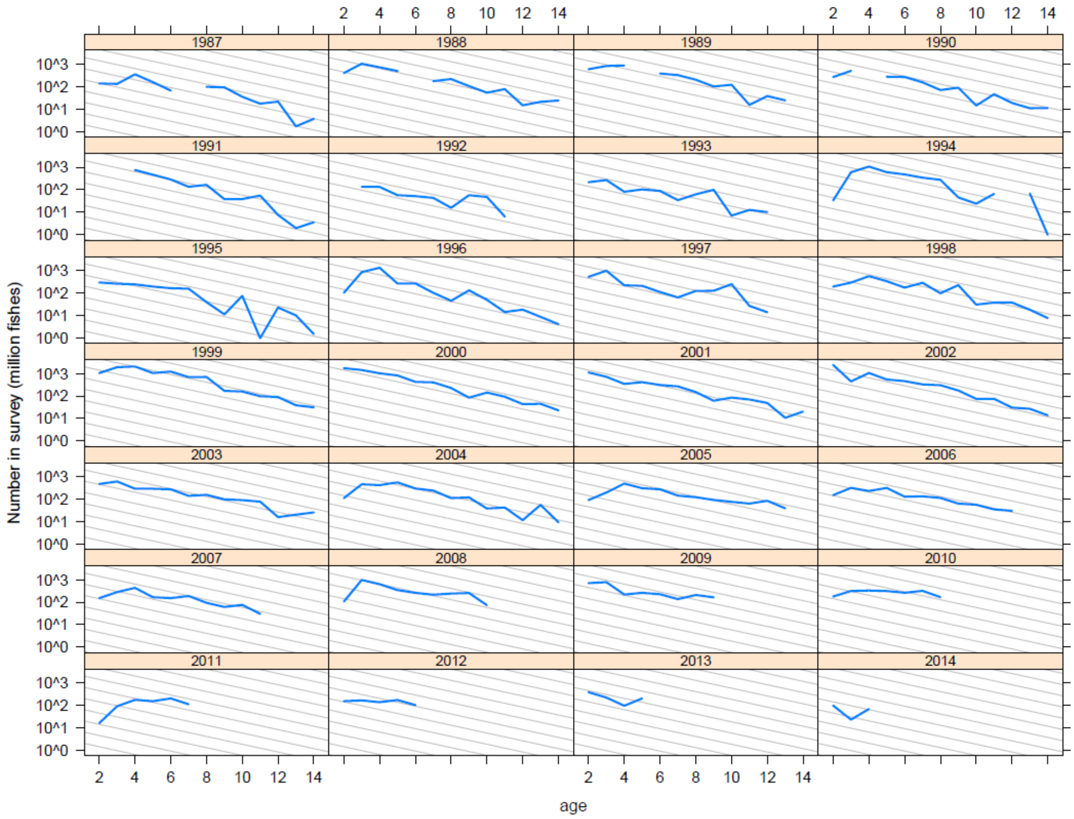

Figure 11.3.1.2. Icelandic summer spawning herring. Catch curves (log_2 of indices) from survey data by year classes 1987–2014. Grey lines correspond to Z = 0.4.

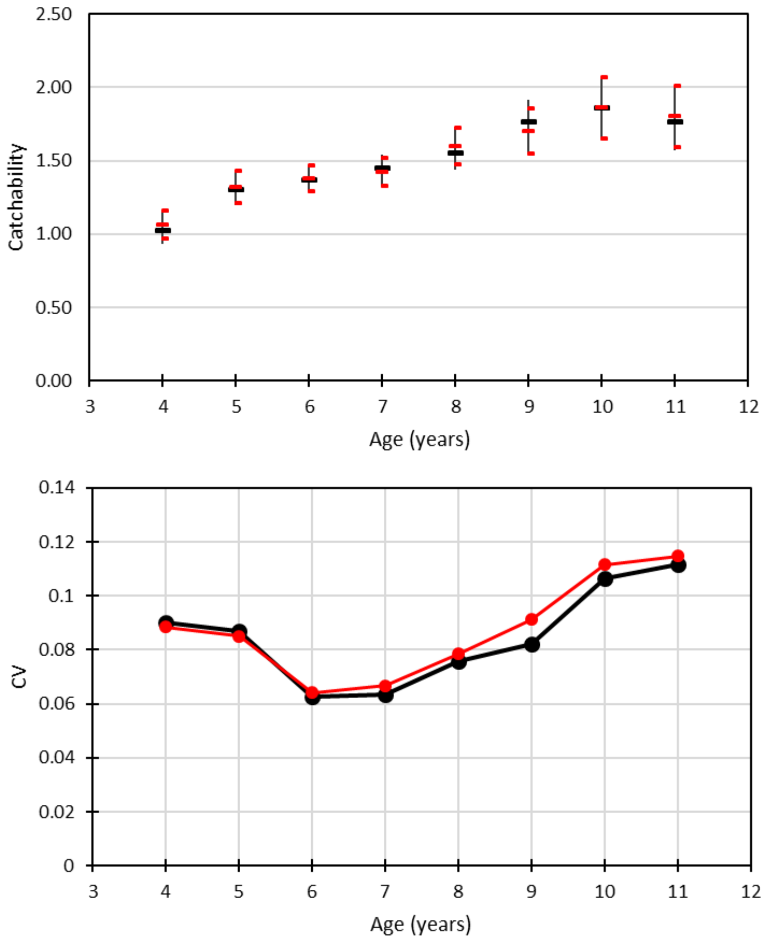

Figure 11.3.2.1. Icelandic summer-spawning herring. The catchability (± 2 SE; Upper graph) and its CV (lower graph) for the acoustic surveys used in the final Adapt run in 2019 (1987–2018) compare to the assessment in 2018 (red lines).

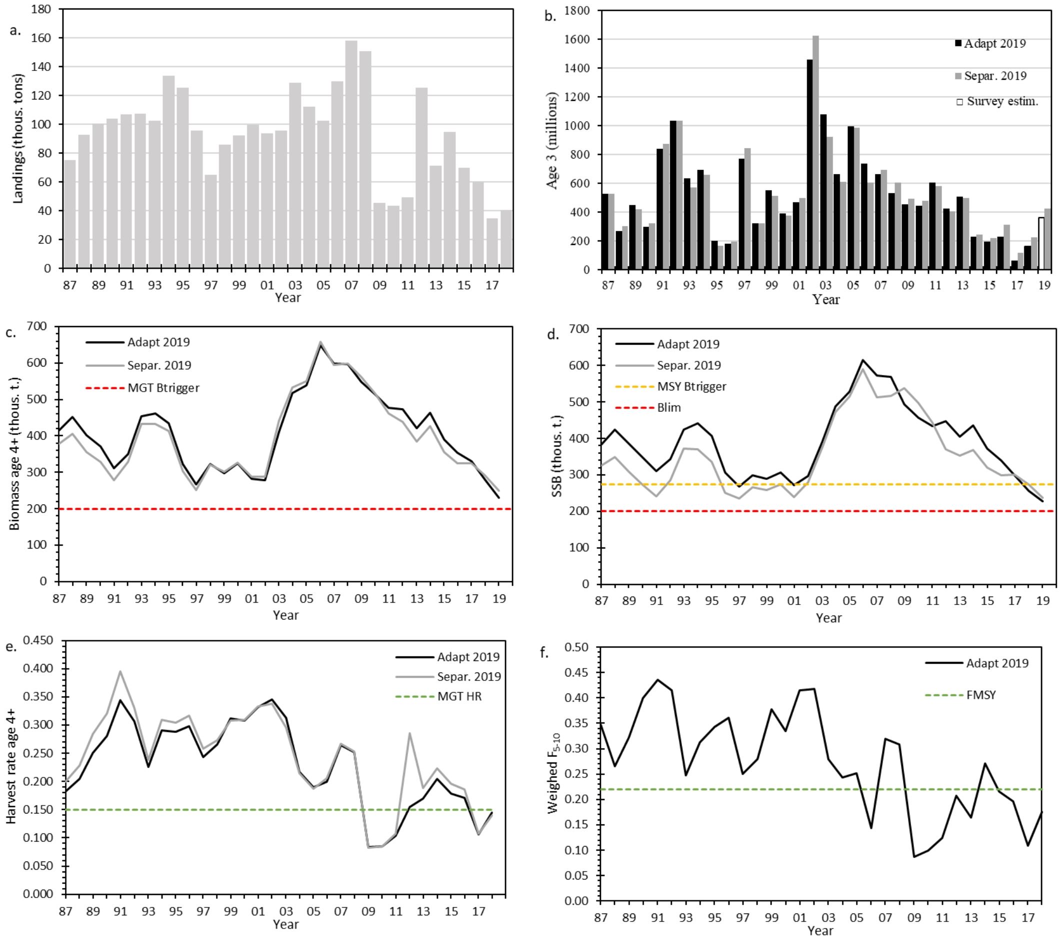

Figure 11.3.2.2. Icelandic summer-spawning herring. Comparisons of the final NFT-Adapt run in 2019 and a run from Separable model (Muppet) in 2018 concerning (a) landings, (b) number at age-3 (recruitment), (c) biomass of age 4+ (reference biomass), (d) SSB, (e) harvest rate of the reference biomass (HR_{MGT} shown), and (f) N-weighed F for age 5–10. Some reference points are also shown. Note that the mass mortality in Kolgrafafjörður in the winter 2012/13 is included in harvest rate for Muppet and not in Adapt. Note that the estimates of number at age 3 in 2017 from Adapt 2019 is not model estimates but derive from survey estimates (i.e. projection from age-1 survey indices in 2017 to age-3 according to Gudmundsdóttir *et al.,* 2007).

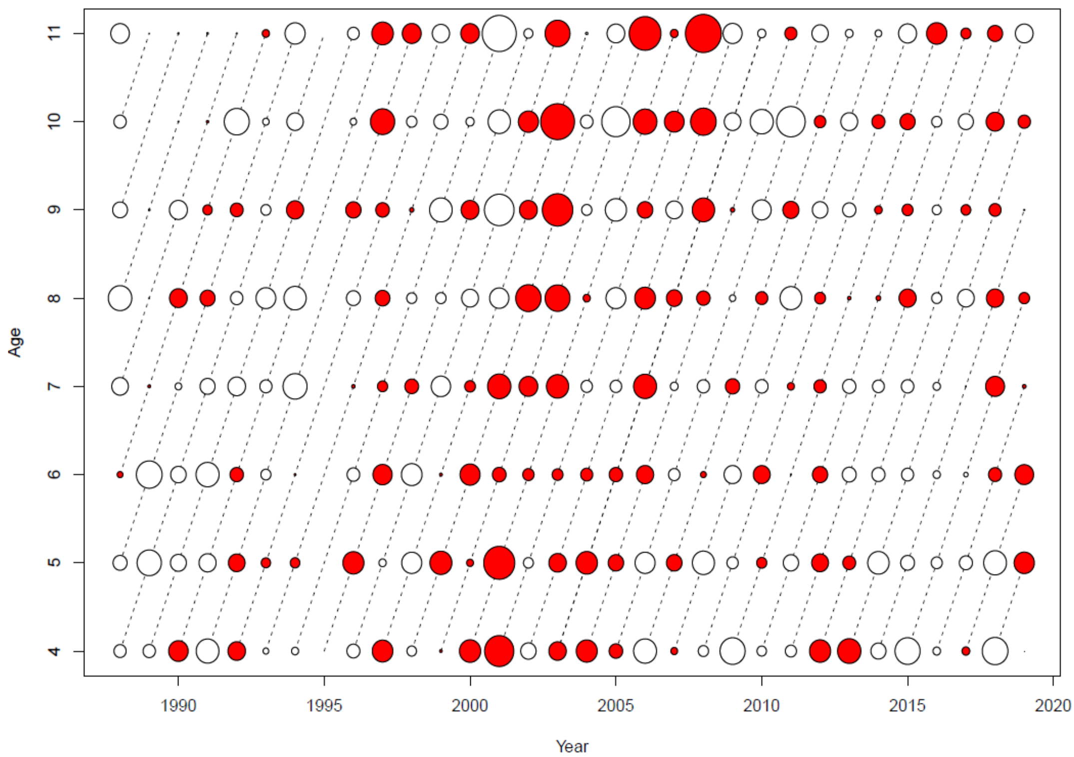

Figure 11.3.2.3. Icelandic summer spawning herring. Residuals of NFT-Adapt run in 2019 from survey observations (moved to 1 January). Filled bubbles are positive (i.e. survey estimates higher than the assessment) and open negative. Max bubble = 1.73.

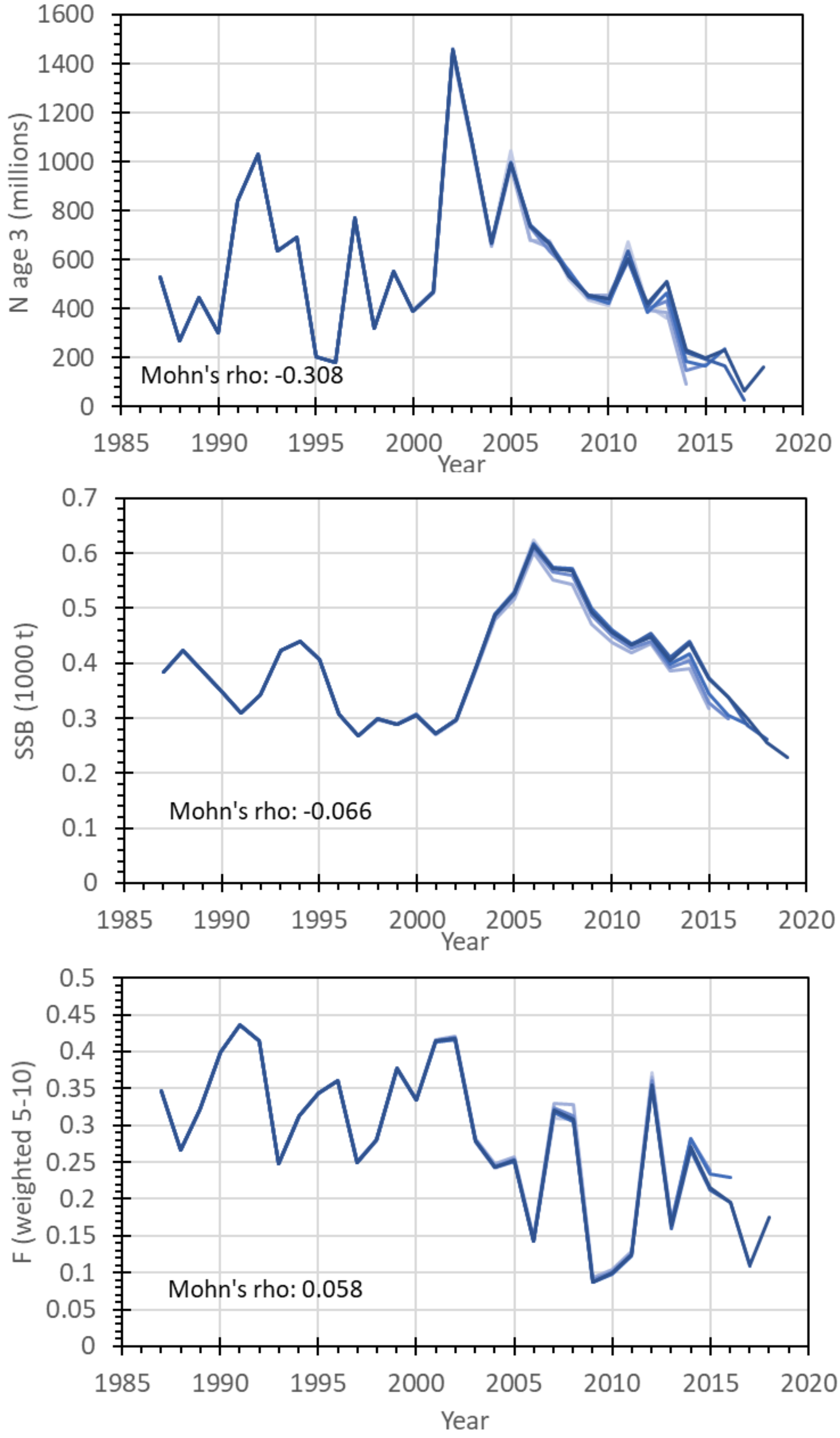

Figure 11.3.2.4. Icelandic summer spawning herring. Six years (2014–2019) retrospective pattern from NFT-Adapt in 2019 in recruitment as number at age 3 (the top panel), spawning stock biomass (middle panel) and N weighted $F_{5–10}$ (lowest panel).

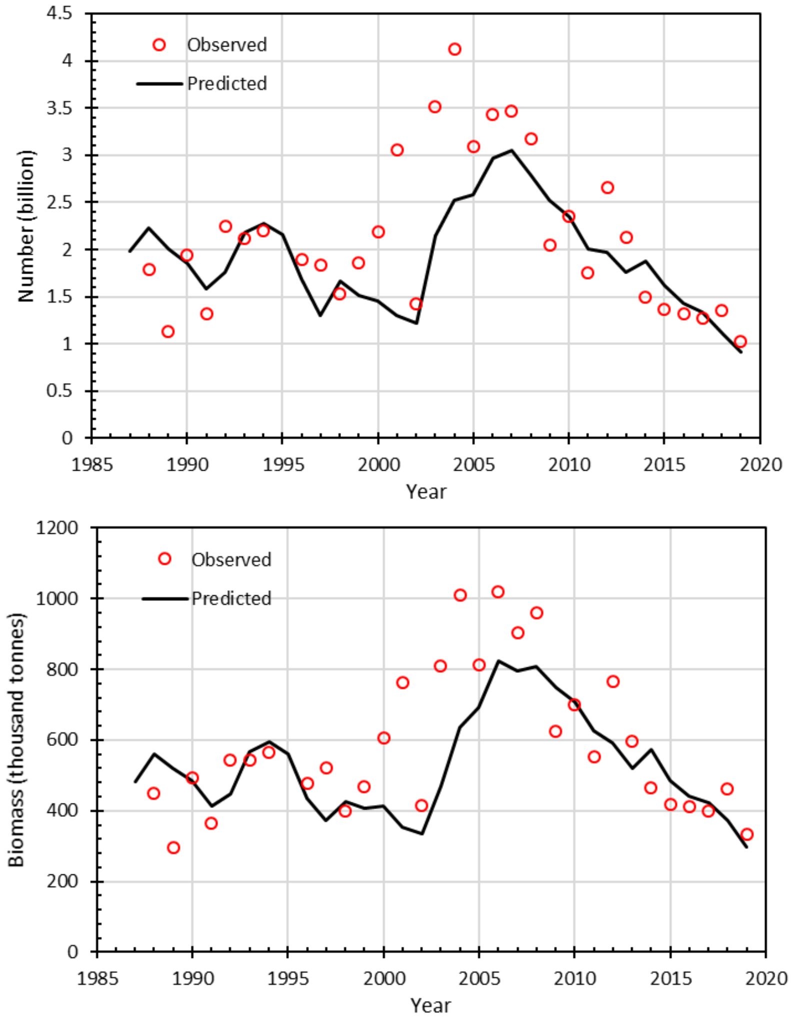

Figure 11.3.2.5. Icelandic summer-spawning herring. Observed versus predicted survey values from NFT-Adapt run in 2019 for ages 4–11 with respect to numbers (upper) and biomass (lower). Note that there was no survey in 1995.

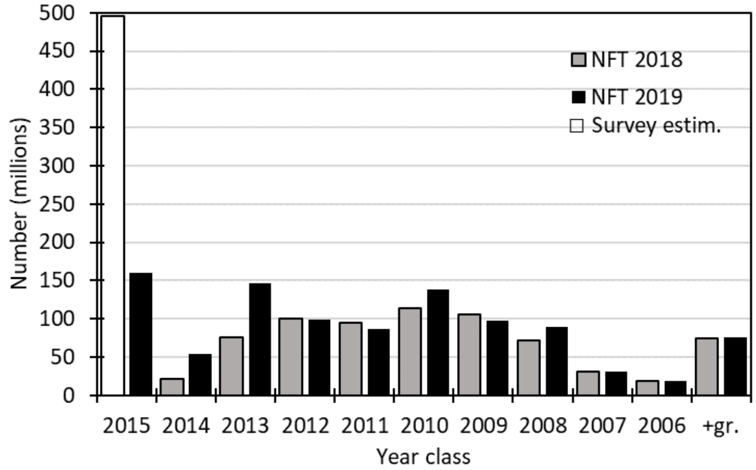

Figure 11.3.2.6. Icelandic summer-spawning herring. Comparison of number-at-age on 1 January 2018 from the final NFT model runs in 2018 and 2019 assessments. Note that the number of the 2015 year from the NFT-2019 is estimated by the model while not from 2018 (based on survey estimate-at- age 1 in 2016) and also the error in the input data in the 2018 assessment (see Section 11.2.3).

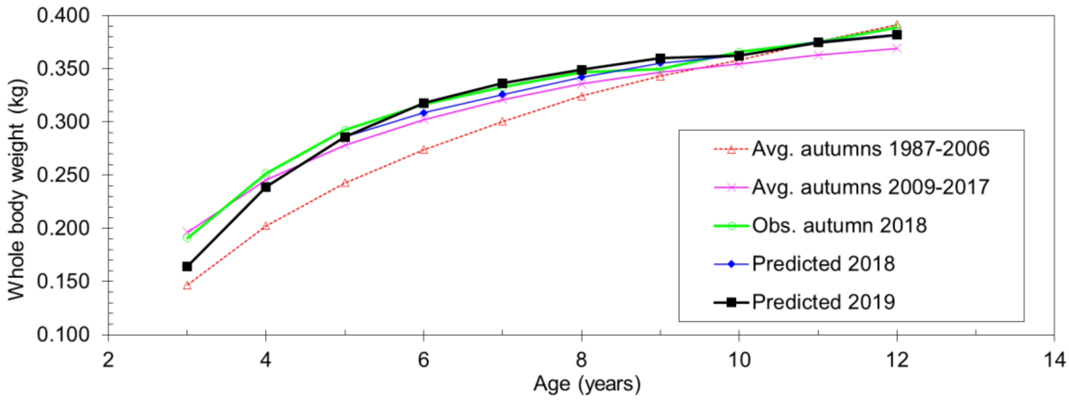

Figure 11.6.1.1. Icelandic summer spawning herring. The mean weight-at-age for age groups 3 to 12 (+ group) in 1987–2006, 2009–2017, in the catches in the autumn 2018, predicted weights for autumn 2018 in the 2018 assessment (ICES, 2018) and finally predicted weights for the autumn 2019 from the weights in 2018, which was used in the stock prognosis.

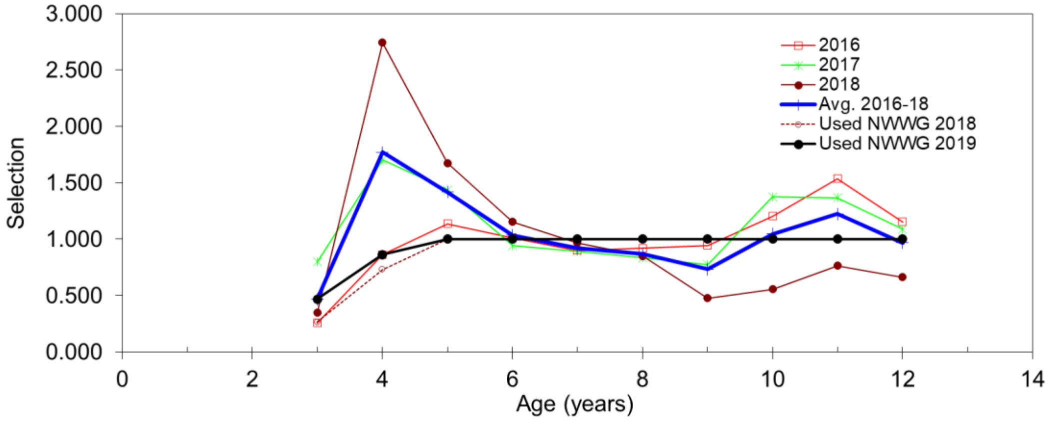

Figure 11.6.1.2. Icelandic summer spawning herring. Estimate of selection pattern ($F_{age}/F_{weighed\ mean\ 5–10}$) in the fishery in the stock prognosis for age groups 3 to 12 (+ group) on basis of the Fs in 2016 to 2018, the average over these three years (used for the prognosis according to the Stock Annex), the selection used in 2018 (ICES, 2018), and the selection used in the prognosis 2019 (deviated from the Stock Annex for age 4 and set equal to 2016 instead of average across 2016–2018).

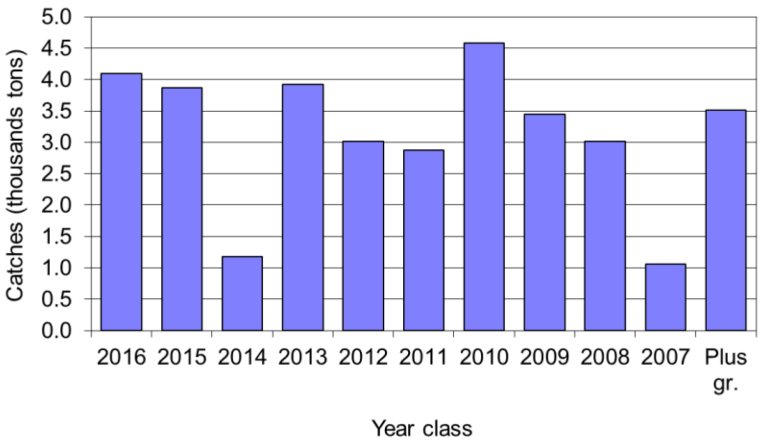

Figure 11.6.2.1. Icelandic summer spawning herring. The predicted biomass contribution of the different year classes to the catches in the fishing season 2019/2020 (total catch of 34 572 tonnes).